

7-12-2013

# Floristics and above-ground biomass (AGB) in Peatlands in Peruvian Lowland Amazonia, Loreto – Peru

Elvis Valderrama

*University of Missouri-St. Louis*

Follow this and additional works at: <http://irl.umsl.edu/thesis>

---

## Recommended Citation

Valderrama, Elvis, "Floristics and above-ground biomass (AGB) in Peatlands in Peruvian Lowland Amazonia, Loreto – Peru" (2013). *Theses*. 26.

<http://irl.umsl.edu/thesis/26>

This Thesis is brought to you for free and open access by the Graduate Works at IRL @ UMSL. It has been accepted for inclusion in Theses by an authorized administrator of IRL @ UMSL. For more information, please contact [marvinh@umsl.edu](mailto:marvinh@umsl.edu).

**Floristics and above-ground biomass (AGB) in Peatlands in Peruvian  
Lowland Amazonia, Loreto – Peru**

Elvis H. Valderrama Sandoval

**A Thesis Submitted to The Graduate School at the University of Missouri-  
St. Louis in partial fulfillment of the requirements for the degree Master of  
Science in Biology**

August 2013

**Advisory Committee**

Peter Stevens, Ph.D.

Chairperson

Amy Zanne, Ph.D.

Ivan Jimenez, Ph.D.

## **Floristics and above-ground biomass (AGB) in Peatlands in Peruvian Lowland Amazonia, Loreto – Peru**

### **Abstract**

Amazonian forests comprise almost 10% of stored carbon (C) in the world's land ecosystems. This C is held both in above-ground biomass (AGB) and in the soil. AGB in an individual plant depends on plant size, often measured in trees as height (H) and diameter (D), and the density of plant tissues, often approximated in trees by wood density (WD). Soil C storage depends on the balance between inputs from AGB due to mortality and senescence and outputs due to decay and erosion. Peatlands, wetlands recently described in northern Peruvian Amazonia, show unusually high rates of soil C accumulation. For these habitats information on C budget contributions from peatland plants is unavailable. In this study I estimated AGB in various peatlands of northern Peruvian Amazonia, and asked why some of these peatlands store more AGB than others. I first set out to estimate the relative contribution of inter- and intra-specific variation to variation in AGB among individual peatland trees. I found that 80% of the variation in AGB among individual trees was due to inter-specific variation. Then I assessed the extent to which the three traits that determine AGB (i.e., D, H and WD) contribute to inter- and intra-specific variation in AGB among peatland trees. I found variation in D and the interaction between D and H contributed most to inter- and intra-specific variation in AGB among trees. Last, I estimated the extent to which variation in AGB among peatland locations was due to variation in species composition, stem density and intra-specific variation in AGB. I found that species composition and intra-specific variation, but not stem density, explained nearly equal amounts of variation in AGB among peatland locations. In summary, detailed knowledge of tree size can provide good estimates of species level biomass estimates in the peatlands of northern Peruvian Amazonia. Additionally, what species are present, as well as how their biomass varies (intra-specifically) from site to site drives AGB variation among peatland locations.

## INTRODUCTION

Accumulation of carbon dioxide (CO<sub>2</sub>) and methane (CH<sub>4</sub>) in the atmosphere are contributing to global climate change (Sha et al., 2011). An understanding of important habitats providing carbon (C) sinks can help in setting conservation priorities to mitigate further climate shifts. Undisturbed Amazonian forests account for approximately 10% of the C stored above-ground in the world's land ecosystems (Anderson, 2012) with woody plants providing short to medium-term C sinks (they can live from 40 to 100 years or more; Vieira et al., 2005) and therefore their C content best governs the above-ground C storage and dynamics in the forest.

Peatlands are poorly drained wetlands that accumulate high quantities of partially decomposed plant organic matter. These habitats are known to play a critical role in storing C from senesced material, but under altered conditions, they could release large amounts of CO<sub>2</sub> (Page et al., 2002) and CH<sub>4</sub> (Urbanová et al., 2013) to the atmosphere. Peatlands have recently been described for the north side of Peruvian Amazonia; they represent approximately 14% of the total area estimated for the entire Amazonian basin and C accumulation rates in the soils (26 to 195 gCm<sup>-2</sup>/y) are similar to those in Indonesia and higher than those in boreal zones (Lahteenoja and Roucoux, 2010). While below-ground C storage has already been described for these habitats (Lahteenoja et al., 2009a; Lahteenoja and Page, 2011); information on C budget contributions from living plants above-ground is not currently available. Here I focus on live standing above-ground biomass (AGB) in the peatlands of northern Peruvian Amazonia, and particularly on three non-exclusive, general, working hypotheses that may explain spatial variation in AGB among peatland locations.

The “floristic composition” hypothesis proposes that variation in AGB across sites results from differences in species composition (e.g. differences in the presence of species and their relative abundance) across sites, because different species are characterized by contrasting AGB values. For example, variation in AGB among eastern and western Amazonian sites was found to be related to variation in the abundance of plant families characterized by relatively

high AGB values, such as Lecythidaceae and Sapotaceae, and others with relatively low AGB values, such as Arecaceae and Myristicaceae (Baker et al. 2004). Variation in environmental conditions (e.g. soil richness and exposure to flooding) and stochastic events across peatland locations (Lahteenoja et al., 2009b) may lead to differences in the composition of species assemblages; and these differences may result in variation in AGB across peatland locations. Some peatlands can be floristically similar to seasonal flooded forests and others to palm forests where *Mauritia flexuosa* (Arecaceae) is common (Lahteenoja et al., 2009a). This variation is thought to influence spatial variation in AGB across sites (Baraloto et al., 2011, Ward et al. 2009). For instance, given that palms are different from other angiosperm trees in growth and tissue structure, the variation in their frequency across sites also may drive variation in biomass estimates. In an exploratory work, Honorio et al. (2009) observed that in forests where the proportion of palms per area was high the total AGB was low.

The “stem density” hypothesis posits that variation in AGB across sites results from site differences in stem density, because increasing the number of stems per area increases AGB. A positive relationship between stem density and AGB has been documented across terra firme forests (Rutishauser et al., 2010). Although such a relationship might seem intuitive, it may not occur when other factors override any potential effect of stem density on AGB. For example, Hawes et. al., (2012) found no association between stem density and AGB across flooded forests; perhaps because variation in exposure to flooding strongly determined AGB via the size of the trees, rather than via the number of trees.

Last, the “intra-specific variation” hypothesis suggests that variation in AGB across sites results from intra-specific variation in AGB among trees in different sites. This hypothesis assumes that at least some species occur in more than one site, and that AGB varies within species from site to site, as illustrated by the aforementioned study of flooded forests (Hawes et. al., 2012). Thus, in sites where species have high AGB values compared to the norm for that species will have high total AGB; conversely, sites where species have low AGB

values will have a correspondingly low total AGB. Intra-specific variation in plant traits can determine important characteristics of species assemblages (Clark 2010, Messier et al., 2010), but it may not have received as much attention as it deserves (Violle et al., 2012).

In this study I estimated AGB in various peatland locations in northern Peruvian Amazonia, and tested predictions from the three hypotheses above to understand why some of these peatland locations store more AGB than others. To set the tests of these predictions in context, I first estimated the relative contribution of inter- and intra-specific variation to variation in AGB among individual peatland trees. Then I assessed how inter- and intra-specific variation in AGB among peatland trees arises from variation among trees in three traits that determine AGB: height (H) diameter (D) and wood density (WD). Last, I estimated the extent to which variation in AGB among peatland locations was explained by three working hypotheses above.

## **METHODS**

### **Study site**

This project was carried out from May to September 2012 in Western Amazonian Peatlands in Loreto, Peru (Appendix; Figure 1). Peatland habitats are distributed along the Pastaza - Marañón basin and the Amazon River and tributaries. The total area of peatlands in this part of Amazonia is unknown but the area in the Pastaza – Marañón basin is approximately 21,929km<sup>2</sup> (Lahteenoja et al., 2011).

For this study, I chose three peatlands (Appendix; Figure 2 a, b, c): San Jorge (SJ, located near the Amazon river, 701852S 9551562W UTM 18 S), Buena Vista (BV, near the Tamshiyacu - Tahuayo river, a tributary of the Amazon, 698280S 9531598W UTM 18 S) and Quistococha (Q, near the Itaya river, another tributary of the Amazon, 686484S 9576261W UTM 18 S). In SJ, the topography is domed and consequently it does not flood. Soils have low nutrients; the forest is swampy and the palm *M. flexuosa* is very common. In both BV and Q, topography is relatively flat, flooding lasts from two to three months

yearly and the forest is swampy. The two forests differ in soil characteristics and floristic composition. BV has rich soils and *M. flexuosa* is almost absent. Q has poor soils and *M. flexuosa* is very common. Detailed soil and topography information has previously been published (Lahteenoja and Page, 2009; Lahteenoja et. al., 2011).

In these peatlands humans hunt mammals (rodents, anteaters) and birds. At San Jorge, villagers from Nueva Vida and San Jorge and nearby villagers at Quistococha use the peatlands mainly to harvest fruits from female *M. flexuosa* palm trees.

### **Climate**

Climate in this part of Amazonia is aseasonal. Total annual precipitation is between 2400 - 3100 mm; the rainiest months are February - April. Minimum temperatures are 20 - 22°C and maximum temperatures are 29 - 31°C. Relative humidity is 80 - 90% (Marengo, 1998).

### **Sampling methods**

The modified Gentry plot design was used for sampling both floristic composition and AGB (Baraloto et al., 2013) with 10 (10 x 50 m) transects summing to a 0.5 ha plot (Appendix; Figure 3). Eight 0.5 ha plots were located in each peatland and were separated by at least 500 m from one another (coordinates of plot locations in Universal Transverse Mercator coordinate system are listed in Appendix, Table 1).

Within each plot all trees with diameter at breast height (DBH)  $\geq$  10 cm were recorded, and their diameter and height measured. Specimen collections were identified to species or separated into morphospecies in the field (e.g. *Pouteria* sp. 1, Myrtaceae sp. 2); at least one voucher specimen per species/morphospecies was preserved for identification at the herbarium. Height was estimated visually to the top of the crown. Trunk core samples were obtained for a subset of individuals following Chave's protocol (2005). Trunk samples were extracted at breast height (1.3 m) for the first 5 individuals (as available) of each species/morphospecies encountered within each plot. It was impossible to reach the pith for all trees (25% lacked pith in cores) as: 1.

increment borers were smaller than the radius of some trees, and 2. some trunks had hollow centers. Length of wood cores varied according to tree diameter from 4 cm - >20 cm. Samples were put into plastic drinking straws, sealed, labeled and stored for later processing.

### **Laboratory work**

Species were identified at the Herbarium Amazonense (Herbarium AMAZ) at the Universidad Nacional de la Amazonia Peruana (UNAP) in Iquitos, Peru. Family names followed the APG III (2009) classification. Voucher specimens were deposited following the series Valderrama, E. et al. 1157-1219.

Wood density estimates (oven dry mass/fresh volume) were obtained at the Facultad de Ingenieria Forestal at UNAP. To be consistent across cores, the core's pith section of all samples was excluded. Core samples were broken into one to eight segments depending on the DBH with each segment 2.5 cm in length. Segments were placed in water within a labeled drinking straw and left for 72 hours to hydrate. Fresh volume (cm<sup>3</sup>) was estimated using the water displacement method (Chave, 2005; Osazuwa-Peters and Zanne, 2012). To determine dry mass, samples were oven dried during three contiguous cycles, each one of 72 hours to ensure a constant mass. Wood density values were estimated for each segment, and mean values across segments were determined for trees with more than one segment.

### **Above-ground biomass estimation**

Above-ground biomass (AGB) was calculated for each tree in the 0.5 ha plots as:

$$AGB = 0.0509 WD \cdot DBH^2 \cdot H \quad (\text{equation 1}),$$

where AGB is above-ground biomass in kg, WD is wood density in g/cm<sup>3</sup>, DBH is diameter at breast height in cm, and H is height in m. Equation 1 is an allometric equation, derived from direct measurements of harvested trees (Chave et al. 2005). The coefficient 0.0509 is an empirical parameter, sometimes called a "form factor" in the forestry literature, used to fit the model to the data. Wood

density was not measured on every tree recorded in all 0.5 ha plots (see above); thus the calculated mean wood density per species per plot was assigned to all individuals of that species within the plot. AGB values were summed across individual trees to obtain total AGB per plot. To facilitate comparison with other studies, values of total AGB per plot were transformed to mega grams per hectare ( $\text{Mg ha}^{-1}$ ).

## Data analyses

***Contribution of inter- and intraspecific variation to AGB variation among peatland trees*** – A variance component analysis that incorporates a random effect for the response variable was used to estimate the amount of total variance in AGB among individual peatland trees that was attributable to variance across and within species.

***Contribution of three plant traits to inter- and intraspecific variation in AGB among peatland trees*** – According to equation 1 (above), AGB is determined by three plant traits: height (H), diameter (DBH) and wood density (WD). I examined how much the variance in each of these traits, and the covariances among traits, contributed to inter- and intra-specific variation in AGB among peatland trees. To do so by way of a simple additive analysis, I first expressed equation 1 in a logarithmic scale:

$$\log(AGB) = \log(0.0509) + \log(WD) + 2 \cdot \log(DBH) + \log(H) \quad (\text{equation 2}),$$

According to well-known properties of the variance of sums (Adler, 1998), the contribution of the variance in each trait, and the covariances among traits, to variance in AGB on a logarithmic scale is:

$$\begin{aligned} \text{Var}[\log(AGB)] = & \text{Var}[\log(WD)] + 4 \cdot \text{Var}[\log(DBH)] + \text{Var}[\log(H)] + \\ & 4 \cdot \text{Cov}[WD, DBH] + 2 \cdot \text{Cov}[WD, H] + 4 \cdot \text{Cov}[DBH, H] \end{aligned} \quad (\text{equation 3}),$$

where  $\text{Var}[\log(AGB)]$  is variance in AGB on a logarithmic scale,  $\text{Var}[\log(WD)]$  variance in wood density on a logarithmic scale,  $\text{Var}[\log(DBH)]$  is variance in

diameter at breast height on a logarithmic scale,  $Var[\log(H)]$  is variance in height on a logarithmic scale, and the remaining terms are the three respective covariances. I applied equation 3 to examine intra-specific variation in each species for which I had at least three measurements of wood density. In the case of inter-specific variation, mean values of each variable were assigned to each species, such that:

$$Var[\log(\overline{AGB})] = Var[\log(\overline{WD})] + 4 \cdot Var[\log(\overline{DBH})] + Var[\log(\overline{H})] + 4 \cdot Cov[\overline{WD}, \overline{AGB}] + 2 \cdot Cov[\overline{WD}, \overline{H}] + 4 \cdot Cov[\overline{DBH}, \overline{H}] \quad (\text{equation 4}),$$

where  $Var[\log(\overline{AGB})]$  is the variance in mean AGB across species on a logarithmic scale,  $Var[\log(\overline{WD})]$  is the variance in mean wood density across species on a logarithmic scale,  $Var[\log(\overline{DBH})]$  is variance in mean diameter at breast height across species on a logarithmic scale,  $Var[\log(\overline{H})]$  is variance in mean height across species on a logarithmic scale, and the remaining terms are the three respective covariances. Mean values for each species were calculated based only on trees whose wood density was measured.

***Testing predictions from three hypotheses about variation in AGB among peatland locations*** – I used one or more simple linear regression models to represent each of the three working hypotheses proposed to explain variation in AGB among peatland sites. For each hypothesis I derived predictions in terms of the statistical significance of regression coefficients and, in some cases, the sign of those coefficients (Table 1).

The floristic composition hypothesis was represented by four regression models. In models 1 and 2, the explanatory variables (NMDS1 and NMDS2) represent the overall species composition. For this, I ran a Non-metric multidimensional scaling analysis (NMDS) to determine the similarity in species composition between plots. I used the Bray-Curtis dissimilarity index (this index takes into account presence and absence of species and also the relative abundance of them). I set the maximum number of random starts at 50 in search

of a stable solution and number of dimensions to three. The stress value obtained in the ordination was 0.0756. The two sets of scores per plot (NMDS1 and NMDS2; for plot scores in each ordination see Appendix; Table 2) were extracted from the NMDS ordination and were included as explanatory variables. The floristic composition hypothesis predicts that the regression coefficients for NMDS1 (in model 1) and NMDS2 (in model 2) are statistically significant (Table 1). No prediction about the sign is possible because the direction of the ordination axes is arbitrary. Models 3 and 4 represent a subset of the overall species composition. In model 3, the proportion of all palms per plot was taken as the explanatory variable and in model 4, only the proportion of *M. flexuosa* was included. The floristic composition hypothesis predicts that the regression coefficients for models 3 and 4 should be statistically significant and negative (Table 1), given previous work suggesting that high relative abundance of palms is associated with low total AGB (Baker et al. 2004, Honorio et al., 2009).

The stem density hypothesis was represented by a single model in which stems density (number of stems/plot) was the explanatory variable. This hypothesis predicts that the regression coefficient of the model should be statistically significant and positive (Table 1).

The intra-specific variation hypothesis was represented by a single regression model in which mean AGB deviation per plot ( $\overline{AGBdeviation}$ ) was the explanatory variable.  $\overline{AGBdeviation}$  values represent the intra-specific variation in AGB across plots. To estimate this, records of all species were considered across all plots. For each species, AGB mean values per plot and across all plots were estimated. Species mean AGB per plot was subtracted from species mean across all plots. Differences of all species within each single plot were summed and divided by the number of species in the plot. Thus,  $\overline{AGBdeviation}$  for a given plot “*k*” with “*s*” species is:

$$\overline{AGBdeviation}_k = \frac{1}{s} \cdot \sum_{i=1}^s (\overline{AGB}_{ik} - \overline{AGB}_i) \quad (\text{equation 5}),$$

where  $\overline{AGB}_{ik}$  is the mean *AGB* for species *i* in plot *k* and  $\overline{AGB}_i$  is the mean *AGB* for species *i* across all plots. A negative  $\overline{AGBdeviation}$  for a given plot indicates that, on average, species at that plot have *AGB* mean values higher than their respective mean values across all plots. Therefore, the intraspecific variation hypothesis predicts a statistically significant regression negative coefficient for  $\overline{AGBdeviation}$  (Table 1).

**Relative importance of the three hypotheses about variation in *AGB* among peatland locations** – I estimated the relative merit of the three hypotheses of interest using two approaches. First, relative empirical support for regression models representing the three hypotheses was gauged using the Akaike Information Criterion in its corrected form for small sample sizes (AICc; Hurvich and Tsai, 1989). Second, I built an *ad hoc* multiple regression model that contained explanatory variables from simple regression models that were significant and also had low AICc values. Different explanatory variables in this *ad hoc* model represented different hypotheses, and thus the relative importance of each hypothesis represented by the model was estimated by the partial coefficient of determination ( $r^2$ ). The correlation between explanatory variables in the *ad hoc* model was low ( $r^2 = 0.07$ ), so collinearity was not a concern.

**Test of spatial dependence** – to explore if the results obtained from regression models were affected by spatial dependence, potentially yielding spurious results, I conducted Moran I test for spatial autocorrelation in regression residuals (Anselin, 2003; Laurent et al., 2012). I applied this test only to the regression models for which I found empirical support.

**Outlier plot** – Plot 6 at SJ had an unusually high value for total *AGB*. Careful examination revealed no measurement or analysis errors that could explain this particularly high value. Nonetheless, since it may have high leverage on the analyses, I ran two sets of analyses, including and excluding this plot.

## RESULTS

**Peatland species assemblages** – One hundred thirty eight species and morphospecies were identified across the 24 plots with 113 identified to species (82.61%), 18 morphospecies identified to genus (12.32%) and 3 to family; 4 remained undetermined. Identified taxa were grouped into 94 genera and 42 families (Appendix, Table 3). The clades with the highest species richness were Fabaceae and *Inga* within Fabaceae (Appendix, Table 4A, 4B). Across the 24 plots, I located 7831 individuals, and *M. flexuosa* in the Arecaceae had by far the greatest number of individual stems (Appendix, Table 5A, 5B, 5C). The total number of species per plot ranged from 11 - 45 and stem density per plot ranged from 233 - 513 (Appendix, Table 6).

The NMDS1 ordination (non-metric multidimensional scaling) explained the bulk of variation in species composition among peatland plots (Figure 1). This axis separated the different peatland locations in the study. BV was quite different floristically with plots loading on the negative end of the ordination axis. SJ and Q were more similar to one another than to BV; they loaded on the positive end of the first ordination axis. Plots at Q and SJ separated along the second ordination axis with Q loading negatively and SJ loading positively.

Variation in total AGB per plot was substantial, ranging from 68.0 to 189.8 Mg ha<sup>-1</sup> (Appendix, Table 6), with BV having the lowest values and SJ having the highest values. Plot 6 at SJ had a remarkably high AGB value in comparison to the other plots. This was due to the presence of many large trees, with high DBH and H values (most had DBH>28 cm and H>16 m). *Platycarpum* sp. nov. was the major contributor since it was the most abundant species (172 out of 355 individuals) and represented 75% of the total AGB in this plot. Further analyses were run with and without this data point included.

**Contribution of inter- and intra-specific variation to AGB variation among peatland trees** – Eighty percent of the total variance in AGB was attributable to differences among species and 20% to intra-specific variability (Appendix, Table 7).

**Contribution of three plant traits to inter- and intra-specific variation in AGB among peatland trees** – The greatest contribution to inter-specific variation in AGB came from variance in diameter and positive covariance between diameter and height (Appendix, Table 8, Figure 2a). On the other hand, for intra-specific variation, for 86 out of 98 species included in the analysis, the greatest contribution to variation in AGB came from variance in diameter. For 79 out of those 86 species, the positive covariance between diameter and height also contributed strongly (Appendix, Table 9, Figure 2b).

**Tests of predictions from three hypotheses about variation in AGB among peatland locations** – In general, I found empirical support for predictions from the floristic composition and intra-specific variation hypotheses, and no support for predictions from the stem density hypothesis. In particular, the regression coefficient for one model representing the floristic composition hypothesis (model 1, where NMDS1 is the explanatory variable representing one axis of the overall species composition across plots) was statistically significant (Table 2, Figure 3a). Likewise, the regression coefficient for the model representing the intra-specific variation hypothesis (model 6, where  $\overline{AGBdeviation}$  is the explanatory variable) was statistically significant and negative (Table 2, Figure 4a). The regression coefficients for all other models were not statistically significant (Table 2).

The overall results change little after excluding the outlier plot 6 at SJ. Empirical support was still restricted to predictions from the floristic composition and intra-specific variation hypothesis, as represented by models 1 and 6, respectively (Table 2, Figures 3b and 4b). Nonetheless, when the outlier was excluded, the regression coefficients of two additional models representing the floristic composition hypothesis were positive and statistically significant, thus opposite in sign to the respective predictions. These were models 3 and 4, in which the proportion of palms and the proportion of *M. flexuosa* palm trees, respectively,  $\overline{AGBdeviation}$  were positively related to AGB (Table 2, Figures 5a, b and 6a, b).

**Relative importance of the three hypotheses about variation in AGB among peatland locations** – For models determined *a priori*, in both sets of models with and without inclusion of the outlier, the lowest AICc corresponded to model 1 (Table 3) meaning that floristic composition was the best predictor of variation in total AGB among peatland plots.

Based on *P* and AICc values, variables from models 1 (NMDS1) and 6  $\overline{AGBdeviation}$  were selected and as the correlation between them was low ( $r^2 = 0.07$ ) they were included in an *ad hoc* model. When retaining the outlier, the overall model was significant (Table 2) explaining 46% of the variation in AGB and with both variables explaining almost the same amount of the variation (partial correlations: NMDS1:  $r^2 = 0.25$ ,  $P = 0.003$ ;  $\overline{AGBdeviation}$  :  $r^2 = 0.22$ ,  $P = 0.008$ ). When the outlier was removed, the same pattern was obtained for the entire model (NMDS1:  $r = 0.27$ ,  $P = 0.002$ ,  $\overline{AGBdeviation}$ :  $r = 0.22$ ,  $P = 0.008$ ). AICc values in both data sets including and excluding the outlier were lower for the multivariate *ad hoc* model than were those for models that were determined *a priori* (Table 3).

**Test of spatial dependence** – Spatial autocorrelation was low for all models (Appendix, table 10) with model 7 having the lowest value.

## DISCUSSION

The variation in above-ground carbon storage in living biomass at the plot level was largely supported by two of the hypotheses (1 and 3) tested in this study. Between and within species differences in AGB largely determined the variation in AGB across study sites, both explaining almost the same amount of variation in the response variable. Additionally, most of the variation in AGB among trees was attributable to the variation among species. Diameter and the positive covariance between D and H had the greatest contribution to inter and intra-specific variation in AGB.

**Support for hypotheses about variation in AGB among peatland locations** – I found empirical support for the floristic composition and the intra-

specific variation hypotheses, but not for the stem density hypothesis. The floristic composition and intra-specific variation hypotheses explained similar amounts of the variation in AGB among peatland locations.

The overall variation in floristic composition was a good predictor of the variation in AGB across plots. When taking the top 24 species at extreme ends of the NMDS1 axis (species scores listed in Appendix, Table 12), mean AGB values were significantly different between the two groups (Kruskal-Wallis chi-squared= 7.2963,  $df= 1$ ,  $p= 0.007$ ). Species composition and the differences in construction of those species should underpin variation in AGB. An interesting example is *M. flexuosa*. At the high end of NMDS1 (plots at Q and SJ), this palm is frequent, is in the top ten species for highest mean H and D values, and in most of the plots it constitutes >30% of the total amount of AGB (Appendix, Table 13). Even if the proportion of this species was not the best predictor of the variation in AGB, it constitutes an important element of the overall species composition as its distribution was significantly correlated with NMDS1 ( $Z=2.73$ ,  $p=0.006$ ,  $\tau=0.41$ ) and NMDS2 ( $Z=-2.13$ ,  $p=0.03$ ,  $\tau=-0.321$ ) and consequently a potential driver of the high AGB values at these plots. This is contrary to what was reported for palm forests where this species was also very common (Honorio et al., 2009). One explanation for these differences may be that I also took into account sites having a different floristic composition (plots at BV) to that of palm forests, and these had lower AGB values at the low end of NMDS1.

I found no support for the stem density hypothesis. The amount of biomass varied independently of the total number of individual stems per area across plots. The variation in environmental conditions already reported for these peatlands (Lahteenoja et al., 2009b) e.g., soil richness, exposure to flooding, could have had an effect, over time, on the growth of trees. In previous studies (Hawes et al., 2012), the size of the trees varied in relation to the time of exposure to flooding. I did not measure these variables; however, I noticed that some plots having relatively high number of stems did not have large biomass values since the trees that were there were small in size. The opposite was observed for some plots having few stems. Additionally, it is important to mention

that during the year I did my sampling an atypical increase in the level of flooding took place in Amazonia. This event had an important effect on the standing biomass. Some plots had small to big fallen, but still living trees. These fallen trees were not taken into account in the study, as I only considered standing living trees for my biomass estimations. The inclusion of the fallen trees in the analysis could have altered the results.

Intra-specific variation in AGB across plots was also a strong predictor of plot level AGB with species able to make large plants in particular locations driving high total AGB at a given plot. The variation in the size of the trees could be a response to the variation of the environmental factors across sites already mentioned in the paragraph above. This phenotypic variation can be thought to be related, on the one hand, to phenotypic plasticity (which may also be a response to the variation in the genotypes of individuals) meaning that some species are capable of responding to local conditions and producing large individuals, consequently leading to high AGB values. However these species may be unable to produce large individuals in all conditions. On the other hand, stochastic (e.g. local storms) and non-random (e.g. seed dispersal limitation) processes occurring at different strengths across plots may have also determined, through time, the variation in the size of individuals. If stochastic processes are at work, intra-specific AGB differences across plots may be strongly related to changes in site-level soil characteristics (e.g. nutrient content, drainage, peat thickness) and frequency of flooding.

***Contribution of three plant traits to inter- and intraspecific variation in AGB among peatland trees*** – Variance in diameter and the (positive) covariance between diameter and height were the components contributing most to the variance in AGB within and across species. For both among and within species variation in AGB, D was the main driver of the variation in AGB. As this trait is squared in the equation, any unit shift in D has greater effect on AGB than does a unit shift in H and WD. On the other hand, according to the covariance between variables, the overall variation in volume (in the equation represented by D times H) should be driving the variation in AGB, with D exerting the strongest

control on this variation. Therefore, an overall shift in size of the trees should have a relatively large impact on biomass and thus the carbon stored.

These results may not be directly comparable to those reported in other studies, where D was also determined as having a strong control on the variation in AGB (Chave et al., 2001; Baraloto et al., 2011; Hawes et al. 2012) because of the differences in the numerical scales (logarithmic in this work). However, it is interesting that I was able to recover a similar pattern. This can be considered a first simple approximation of decomposing an AGB equation to determine which traits are drivers of variation in AGB.

**Some general aspects** - Values of total AGB reported in this work were lower than those for other habitats in Amazonia (Baker et al., 2004) but similar to those documented for forests where palms, including *M. flexuosa*, are common (Honorio et al., 2009; Appendix, Table 11).

The high AGB in the outlier plot (plot 6) seems to be due to the presence of trees that are relatively large in size with *Platycarpum* sp. nov contributing more to this given that it was the most common species in this plot. Species of *Platycarpum* tend to occupy habitats having poor soil conditions (King, 1984) and some tend to attain large sizes; for example, *Platycarpum orinocense* is an endemic and common canopy species in white sand forests (Vriesendorp et al., 2006). *Platycarpum* sp. nov. was also reported at Jenaro Herrera in Peruvian lowland Amazonia, another palm forest, where *M. flexuosa* is very common (Davila *com. pers*). While this species is also common in other locations, its biology is still relatively unknown and under study (Davila, *in press*). Other species that also were big and contributed, although to a lesser extent, to the extremely high AGB values in this plot were *M. flexuosa* and *Ficus guianensis*, the first one representing 9.48% of the total AGB and the second having the biggest tree in the plot (D = 56 cm). Many of the species in this plot were also found in other plots. Various biotic and abiotic characteristics at plot 6 in SJ likely allowed individuals of these species to attain higher sizes in D and H than those of the same species in other plots.

## CONCLUSIONS

In this first study of AGB in Peruvian peatlands, I showed that most variance (80%) in AGB among peatland trees is attributable to differences among species; that stem size is a major determinant of variation in AGB among peatland trees, both among and within species; and that variation in total AGB among peatland locations is equally determined by floristic composition and intra-specific variation in AGB. I also showed that *M. flexuosa* is an important element of the overall species composition and is a potential driver of the high AGB values in plots where it occurred.

I sampled a subset of the northern Peruvian peatlands that differed in flooding frequency and nutrient content in the soils. In future work, it would be useful to collect information across a greater breadth of Peruvian peatlands in order to see if patterns observed in this work are maintained. Additionally, a better understanding of variation in AGB across the landscape can be gained by understanding the drivers of species distributions, as well as what leads to phenotypic plasticity in species construction. The results from this study can be used to better parameterize site C budgets beyond just the amount of C stored in the soil. Future studies of C budgets should also include estimates of C stored in fallen and standing dead trees as well as decomposition rates both for woody debris and leaf litter, as well as the soil (Větrovský et al. 2011). Finally, to prioritize sites for forest management, species composition, the presence of *M. flexuosa*, and tree size are among the variables that should be taken into account.

## REFERENCES

- Adler, F. 1998. Modeling the Dynamics of Life: Calculus and Probability for Life Scientists. Brooks/Cole. 1<sup>st</sup> Edition. Mathematics Ser. Pp. 800.
- Anderson, L. O. 2012. Biome-scale forest properties in Amazonia based on field and satellite observations. Remote Sensing 4: 1245-1271.
- Anselin, L. 2003. An Introduction to Spatial Regression Analysis in R. University of Illinois, Urbana-Champaign. <http://sal.agecon.uiuc.edu>

- Baker, T. R., O. L. Phillips, Y. Malhi, S. Almeida, L. Arroyo, A. di Fiore, T. Erwin, T. J. Killeen, S. G. Laurance, W. F. Laurance, S. L. Lewis, J. Lloyd, A. Monteagudo, D. A. Neill, S. Patino, N. C. A. Pitman, J. N. M. Silva and R. Vasquez Martinez. 2004. Variation in wood density determines spatial patterns in Amazonian forest biomass. *Global Change Biology* 10: 545-562.
- Baraloto, C., S. Rabaud, Q. Molto, L. Blanc, C. Fortunel, B. Hérault, B., N. Dávila, I. Mezones, M. Rios. E. Valderrama and P. Fine. 2011. Disentangling stand and environmental correlates of aboveground biomass in Amazonian forests. *Global Change Biology* 17: 2677-2688.
- Baraloto, C., Q. Molto, S. Rabaud, B. Hérault, R. Valencia, L. Blanc, P. Fine and J. Thompson. 2013. Rapid simultaneous estimation of aboveground biomass and tree diversity across neotropical forests: A comparison of field inventory methods. *Biotropica* 45: 288-298.
- Chave, J., B. Riera and M. A. Dubois. 2001. Estimation of biomass in a neotropical forest of French Guiana: Spatial and temporal variability. *Journal of Tropical Ecology* 10: 545-562.
- Chave, J. 2005. Wood density measurement protocol - Measuring wood density for tropical forest trees, A field manual for the CTFS sites.
- Chave J., C. Andalo, S. Brown, M. A. Cairns, J. Q. Chambers, D. Eamus, H. Folster, F. Fromard, N. Higuchi, T. Kira, J.-P. Lescure, B. W. Nelson, H. Ogawa, H. Puig, B. Riera and T. Yamakura. 2005. Tree allometry and improved estimation of carbon stocks and balance in tropical forests. *Oecologia* 145: 87-99.
- Clark, S. J. 2010. Individuals and the Variation Needed for High Species Diversity in Forest Trees. *Science* 327: 1129 – 1132.
- Davila, N. *In press*. Sistemática e Biogeografía de Henriquezieae (Rubiaceae).
- Hawes, J.E., C. A. Peres, L.B. Riley, and L. L. Hess. 2012. Landscape-scale variation in structure and biomass of Amazonian seasonally flooded and unflooded forests. *Forest Ecology and Management* 281: 163 -176

- Hurvich, C. M. and C.-L. Tsai. 1989. Regression and time series model selection in small samples. *Biometrika* 76: 297-307.
- Honorio, E., M. Corrales, A. Vásquez, J. Irarica, N. Saavedra and H. Vásquez. 2009. Evaluación de almacenamiento y secuestro de carbono en aguajales de Loreto, Peru. Memoria institucional, Loreto, Ucayali, San Martín, Madre de Dios, Huánuco (Tingo María) y Amazonas – Instituto de Investigaciones de la Amazonia Peruana. Pp. 31-32.
- King, R.G. 1984. Flora Neotropica, Gleasonia, Henriquezia, and Platycarpum (Rubiaceae). Monograph 39: Pp. 134.
- Lähteenoja, O., K. Ruokolainen, L. Schulman and M. Oinonen. 2009a. Amazonian peatlands: An ignored C sink and potential source. *Global Change Biology* 15: 2311-2320.
- Lähteenoja, O., K. Ruokolainen, L. Schulman and J. Alvarez. 2009b. Amazonian floodplains harbour minerotrophic and ombrotrophic peatlands. *Catena* 79: 140-145.
- Lähteenoja, O. and K. H. Roucoux. 2010. Inception, history and development of peatlands in the Amazon basin. *PAGES news* 18(1): 27-29.
- Lähteenoja, O. and S. Page. 2011. High diversity of tropical peatland ecosystem types in the Pastaza-Maranon basin, Peruvian Amazonia. *Journal of Geophysical Research* 116: 148-227.
- Lähteenoja, O., Y. Rojas Reategui, M. Rasanen, D. Del Castillo Torres, M. Oinonen and S. Page. 2011. The large Amazonian peatland carbon sink in the subsiding Pastaza-Maranon foreland basin, Peru. *Global Change Biology* 18: 164-178.
- Lauren, T., A. Ruiz-Gazen and C. Thomas-Agnan. 2012. Some examples of use of the GeoXp package (version 1.5.0). Toulouse School of Economics. [http://cran.rproject.org/web/packages/GeoXp/vignettes/presentation\\_geoxp.pdf](http://cran.rproject.org/web/packages/GeoXp/vignettes/presentation_geoxp.pdf)
- Marengo, J. A. 1998. Climatología de la zona de Iquitos, Perú. In *Geoecología y Desarrollo Amazónico: Estudio Integrado en la Zona de Iquitos, Perú*,

- Ann. Univ. Turkuensis, Ser. A II 114, edited by R. Kalliola and S. Flores Paitán. Univ. of Turku, Turku, Finland. Pp. 35–57.
- Messier, J., B. J. McGill and M. J. Lechowicz. 2010. How do traits vary across ecological scales? A case for trait-based ecology. *Ecology Letters* 13: 838-848.
- Osazuwa-Peter, O., and A. E. Zanne. 2011. Wood density protocol. <http://prometheuswiki.publish.csiro.au/tiki-pagehistory.php?page=Wood%20density%20protocol&preview=11>
- Page, S. E., F. Siegert, J. O. Rieley, H. D. V. Boehm, A. Jaya and S. Limin. 2002. The amount of carbon released from peat and forest fires in Indonesia during 1997. *Nature* 420: 61-65.
- Rutishauser, E., F. Wagner, B. Herault, E. A. Nicolini and L. Blanc. 2010. Contrasting above-ground biomass balance in a Neotropical rain forest. *Journal of Vegetation Science* 21: 672–682.
- Sha, C.-Y., M. Wang, Q. Wang and J.-J. Lu. 2011. Wetland methane and carbon dioxide emission and affecting factors. *Chinese Journal of Ecology* 30: 2072-2079.
- Urbanová, Z., J. Bárta and T. Pícek. 2013. Methane emissions and methanogenic Archaea on pristine, drained and restored mountain peatlands, Central Europe. *Ecosystems* 16: 664-677.
- Větrovský, T., J. Voříšková, J. Šnajdr, J. Gabriel and P. Baldrian. 2011. Ecology of coarse wood decomposition by the saprotrophic fungus *Fomes fomentarius*. *Biodegradation* 22: 709-718.
- Vieira, S., S. Trumbore, P. B. Camargo, D. Selhorst, J. Q. Chambers, N. Higuchi and L.A. Martinelli. 2005. Slow growth rates of Amazonian trees: Consequences for carbon cycling. *Proceedings of the National Academy of Sciences U.S.A.* 102: 18502-18507.
- Violle, C., B. J. Enquist, B. J. McGill, L. Jiang, C. H. Albert, C. Hulshof, V. Jung and J. Messier. 2012. The return of the variance: Intraspecific variability in community ecology. *Trends in Ecology and Evolution* 27: 244-252.

Vriesendorp, C., N. Pitman, J. I. Rojas Moscoso, B. A. Pawlak, L. Rivera Chávez, L. C. Méndez, M. Vela Collantes and P. F. Fasabi Rimachi. 2006. Peru: Matses: Rapid biological inventories 16. The Field Museum, Chicago, IL. Pp. 131.

Ward, S. E., R. D. Bardgett, N. P. McNamara and N. J. Ostle. 2009. Plant functional group identity influences short-term peatland ecosystem carbon flux: Evidence from a plant removal experiment. *Functional Ecology* 23: 454-462.

**Table 1:** Regression models for testing variation in Above-ground biomass.  $\overline{AGBdeviation}_{plot}$  = mean AGB deviation per plot.  $NMDS1_{plot}$  and  $NMDS2_{plot}$  = plot scores that were extracted from the NMDS ordination (the first and second axes) of species composition.  $Ppalms_{plot}$  = proportion of palms per plot.  $Pmauritiam_{plot}$  = proportion of *M. flexuosa* per plot.  $logTsdensity_{plot}$  = logarithm of total number of stems per plot.  $TotalAGB_{plot}$  = total above-ground biomass per plot.

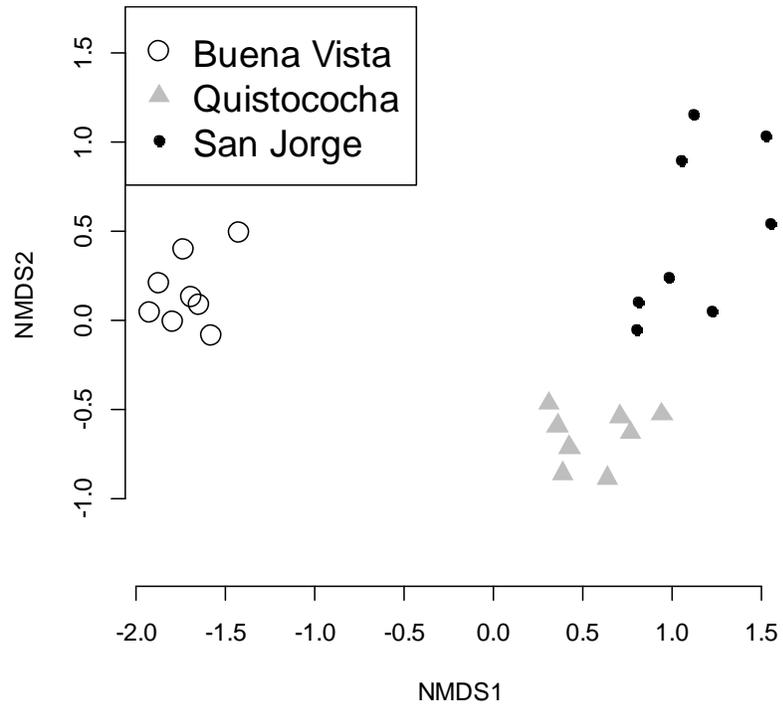
Hypothesis	Linear regression model	Prediction about b
Floristic composition	$TotalAGB_{plot} = a + b NMDS1_{plot}$	Significant
	$TotalAGB_{plot} = a + b NMDS2_{plot}$	Significant
	$TotalAGB_{plot} = a + b Ppalms_{plot}$	Significant -
	$TotalAGB_{plot} = a + b Pmauritiama_{plot}$	Significant -
Stem density	$TotalAGB_{plot} = a + b logTsdensity_{plot}$	Significant +
Intraspecific variation	$TotalAGB_{plot} = a + b \overline{AGBdeviation}_{plot}$	Significant -

**Table 2:** *a priori* and *ad hoc* models predicting plot level aboveground biomass including and excluding the outlier plot 6 at SJ.  $\overline{AGBdeviation}_{plot}$  = mean AGB deviation per plot.  $NMDS1_{plot}$  and  $NMDS2_{plot}$  = plot scores that were extracted from the NMDS ordination (the first and second axes) of species composition.  $Ppalms_{plot}$  = proportion of palms per plot.  $Pmauritia_{plot}$  = proportion of *M. flexuosa* per plot.  $\log Tsdensity_{plot}$  = logarithm of total number of stems per plot.  $TotalAGB_{plot}$  = total aboveground biomass per plot.

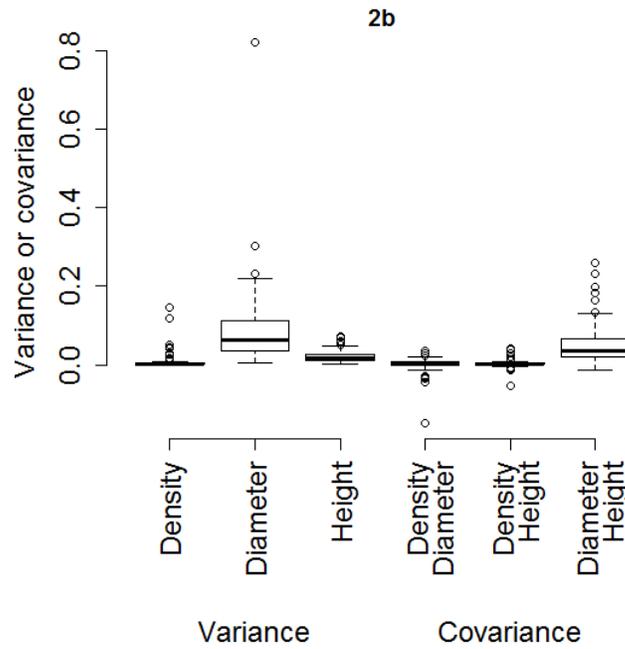
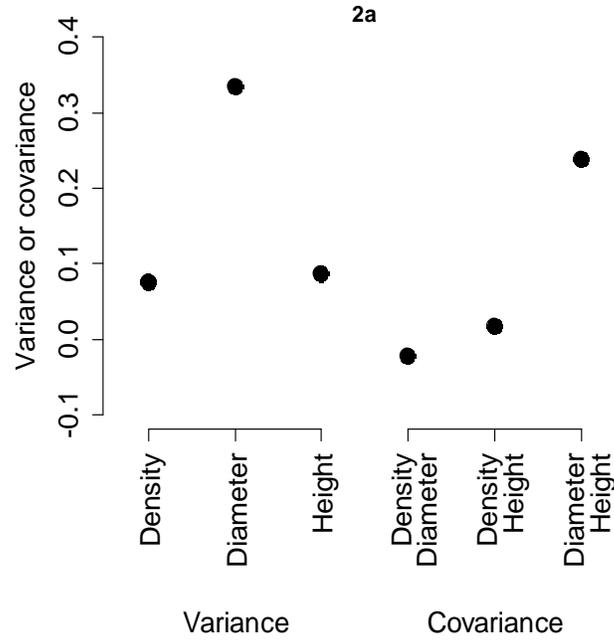
Hypothesis	Linear regression model	Including the outlier					Excluding the outlier				
		<i>F</i>	<i>df</i>	<i>R</i> <sup>2</sup>	<i>B</i>	<i>P</i>	<i>F</i>	<i>df</i>	<i>R</i> <sup>2</sup>	<i>b</i>	<i>P</i>
<b><u>a priori</u></b>											
Floristic composition	1: $TotalAGB_{plot} = a + b NMDS1_{plot}$	7.2	1, 22	0.25	10.4	0.01	7.69	1, 21	0.27	8.25	0.01
	2: $TotalAGB_{plot} = a + b NMDS2_{plot}$	0.009	1, 22	4E-04	0.98	0.9	1.85	1, 21	0.08	-10.38	0.18
	3: $TotalAGB_{plot} = a + b Ppalms_{plot}$	1.09	1, 22	0.05	0.25	0.31	4.69	1, 21	0.18	0.38	0.04
	4: $TotalAGB_{plot} = a + b Pmauritia_{plot}$	1.16	1, 22	0.05	0.39	0.29	4.33	1, 21	0.17	0.54	0.05
Stem density	5: $TotalAGB_{plot} = a + b \log Tsdensity_{plot}$	2.32	1, 22	0.1	0.13	0.14	2.61	1, 21	0.11	0.10	0.12
Intra-specific variation	6: $TotalAGB_{plot} = a + b \overline{AGBdeviation}_{plot}$	4.63	1, 22	0.17	-0.24	0.04	4.05	1, 21	0.16	-0.17	0.05
<b><u>ad hoc</u></b>											
Floristic composition and Intra-specific variation	7: $TotalAGB_{plot} = a + b NMDS1_{plot} + \overline{AGBdeviation}_{plot}$	9.04	2, 21	0.46		0.001	9.67	2, 21	0.49		0.001

**Table 3:** AICc values of *a priori* and *ad hoc* models that explain variation in AGB among peatland plots when including and excluding the outlier.  $\overline{AGBdeviation}_{plot}$  = mean AGB deviation per plot.  $NMDS1_{plot}$  and  $NMDS2_{plot}$  = plot scores that were extracted from the NMDS ordination (the first and second axes) of species composition.  $Ppalms_{plot}$  = proportion of palms per plot.  $Pmaurititia_{plot}$  = proportion of *M. flexuosa* per plot.  $\log Tsdensity_{plot}$  = logarithmic form of total number of stems per plot.  $TotalAGB_{plot}$  = total aboveground biomass per plot.

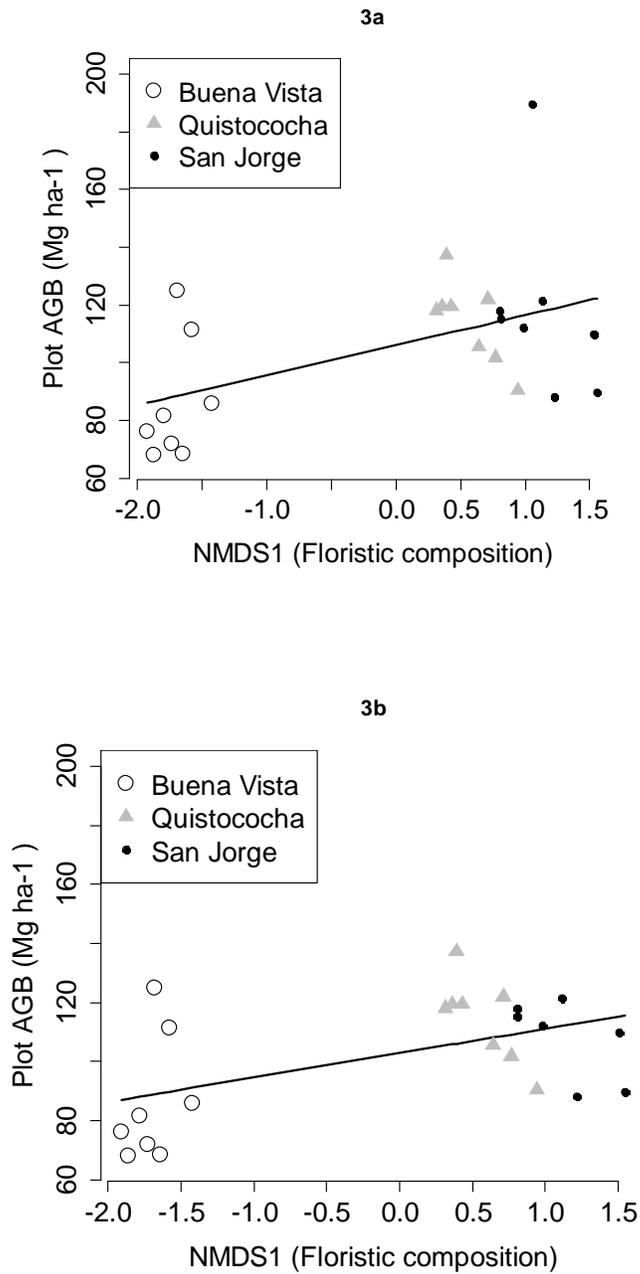
Hypothesis		Linear regression model	AICc
<b><i>a priori</i></b>			
Including the outlier	Floristic composition	1. $TotalAGB_{plot} = a + b NMDS1_{plot}$	224
		3. $TotalAGB_{plot} = a + b Ppalms_{plot}$	229.6
		4. $TotalAGB_{plot} = a + b Pmaurititia_{plot}$	229.6
Intra-specific variation	6. $TotalAGB_{plot} = a + b \overline{AGBdeviation}_{plot}$	226.2	
<b><i>ad hoc</i></b>			
	Floristic composition and intra-specific variation	7. $TotalAGB_{plot} = a + b NMDS1_{plot} + \overline{AGBdeviation}_{plot}$	217.9
<b><i>a priori</i></b>			
Excluding the outlier	Floristic composition	1. $TotalAGB_{plot} = a + b NMDS1_{plot}$	201.68
		3. $TotalAGB_{plot} = a + b Ppalms_{plot}$	204.22
		4. $TotalAGB_{plot} = a + b Pmaurititia_{plot}$	204.55
Intra-specific variation	6. $TotalAGB_{plot} = a + b \overline{AGBdeviation}_{plot}$	204.8	
<b><i>ad hoc</i></b>			
	Floristic composition and intra-specific variation	7. $TotalAGB_{plot} = a + b NMDS1_{plot} + \overline{AGBdeviation}_{plot}$	195.31



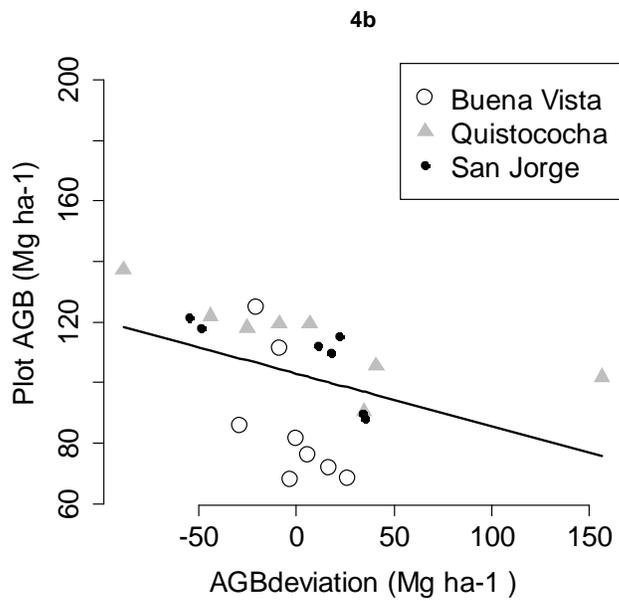
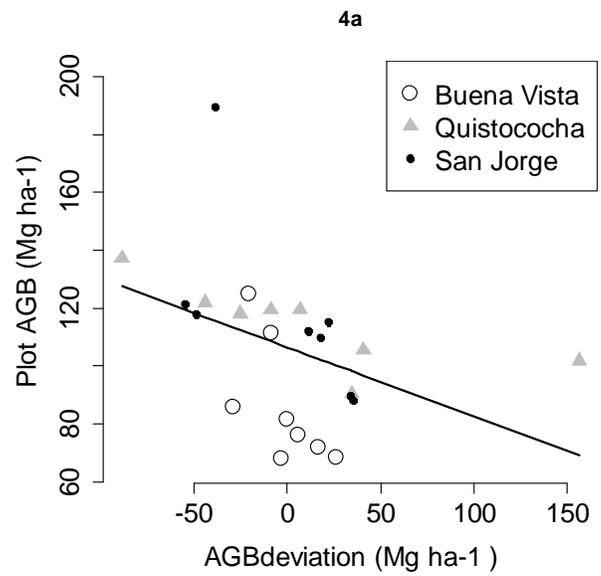
**Figure 1:** NMDS ordination biplot of floristic composition dissimilarities among peatland plots. The stress value of the ordination was 0.075.



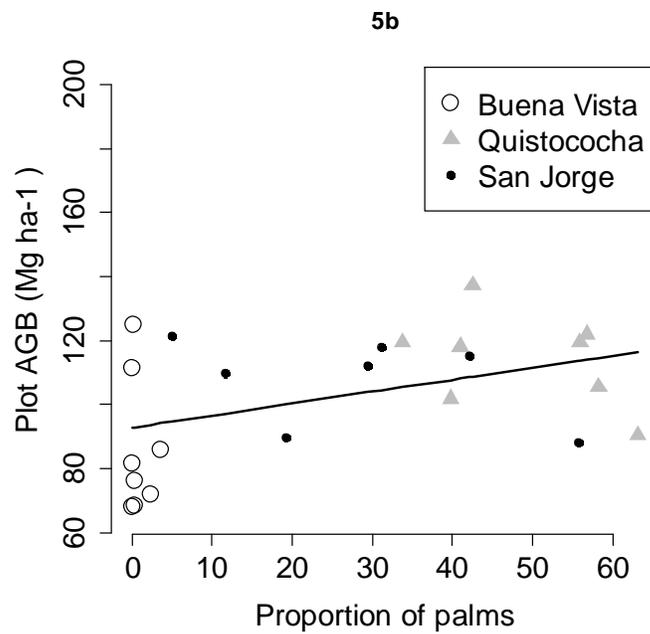
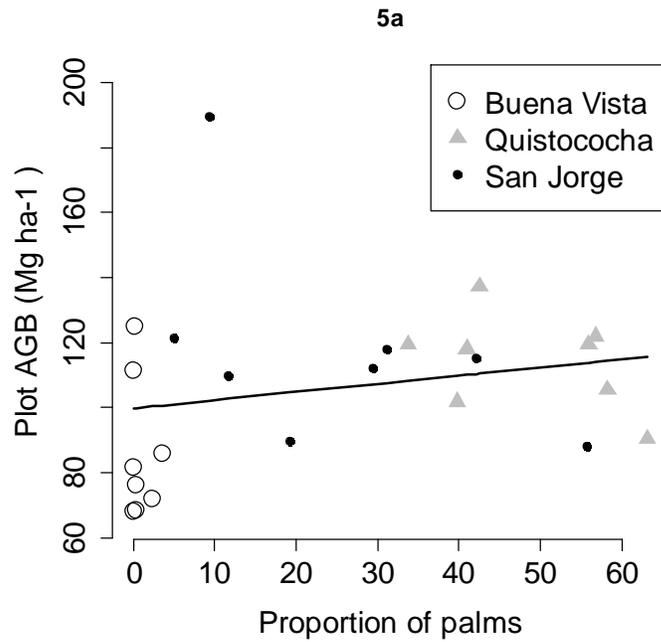
**Figure 2:** Variance and covariance between variables from the equation for estimating AGB. **a** – Inter-specific variation, one single value is shown per each trait. **b** – Intra-specific variation, mean values across the 98 species are shown per each trait, Density = wood density.



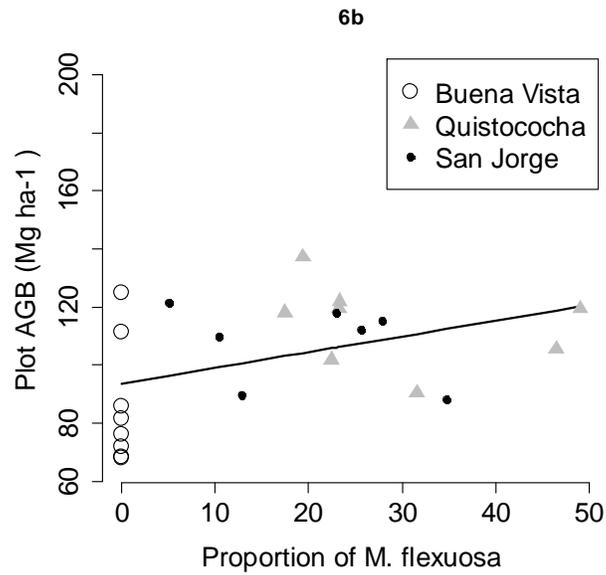
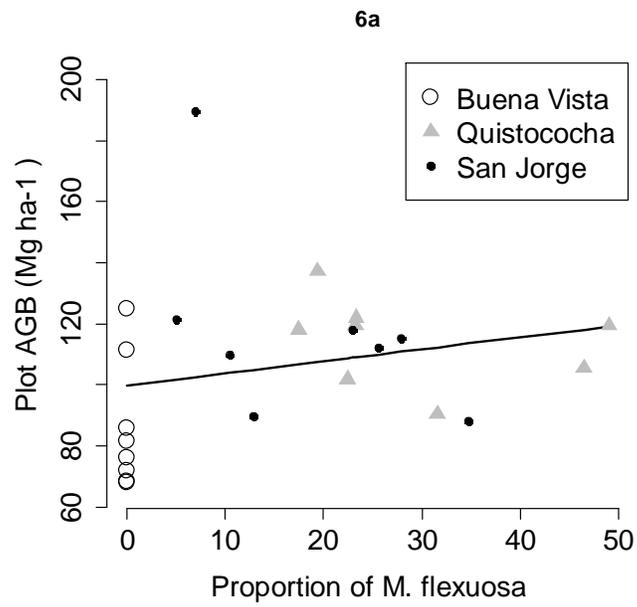
**Figure 3:** Linear regression between total plot AGB and the overall floristic composition (NMDS1). **a** - When the outlier (Plot 6 at SJ) is included, **b** - When the outlier is excluded.



**Figure 4:** Linear regression between total plot AGB and intra-specific variation in AGB ( $\overline{AGBdeviation}$ ). **a** - When the outlier (Plot 6 at SJ) is included, **b** - When the outlier is excluded.



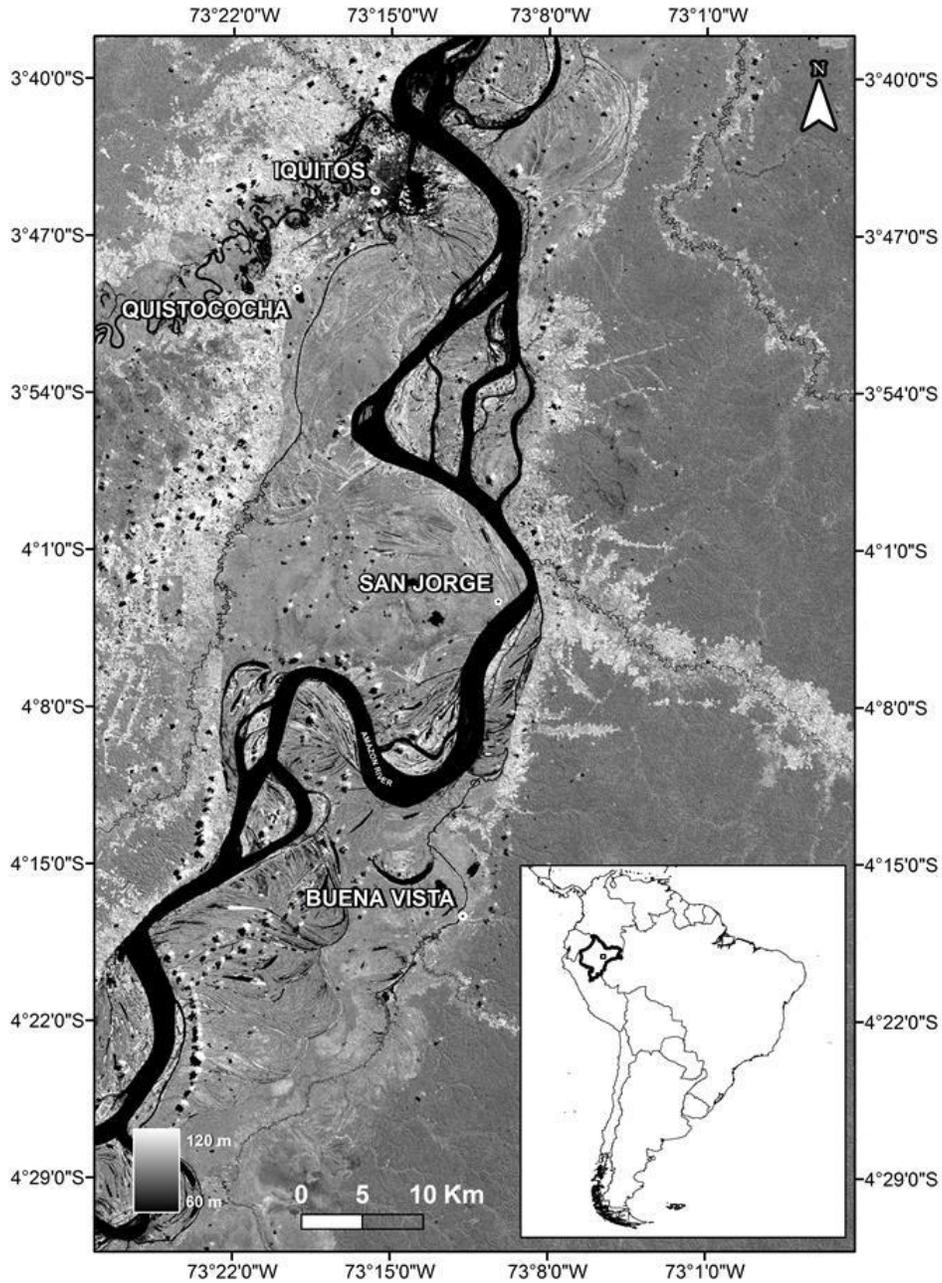
**Figure 5:** Linear regression between total plot AGB and proportion of palms. **a** - When the outlier (Plot 6 at SJ) is included, **b** - When the outlier is excluded.



**Figure 6:** Linear regression between total plot AGB and proportion of *M. flexuosa*. **a** - When the outlier (Plot 6 at SJ) is included, **b** - When the outlier is excluded.

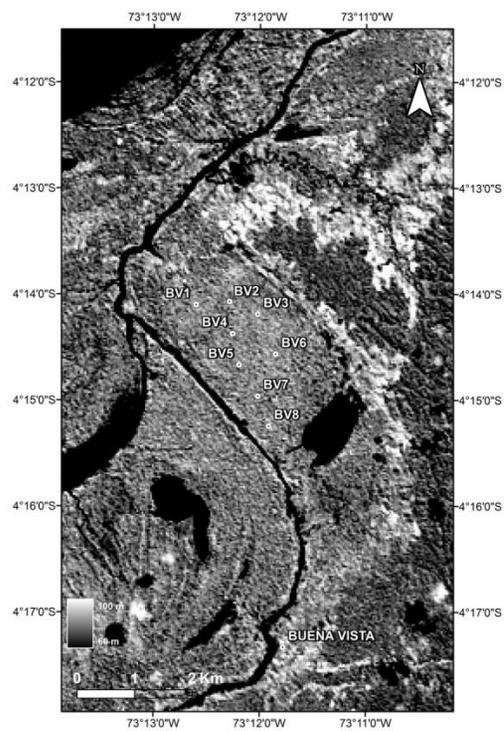
## APENDIX

**Figure 01:** Map showing the location of study sites in Northern Peruvian Amazonia. Study sites are distributed along the Amazon River. The closest city to these is Iquitos (main city in this part of Amazonia)

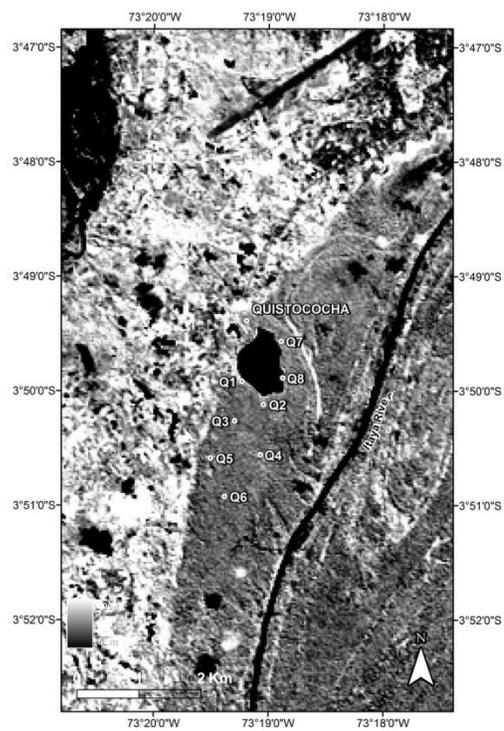


**Figure 02:** Location of plots in Peru in each peatland of study. (a) Buena Vista, (b) Quistococha, (c) San Jorge.

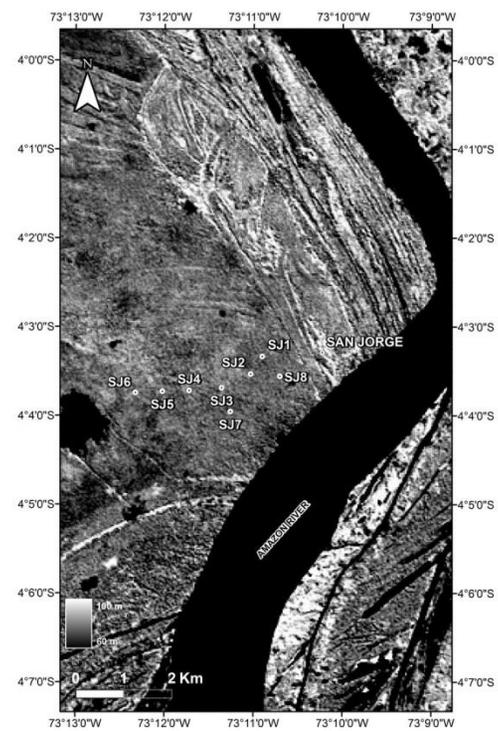
**a**



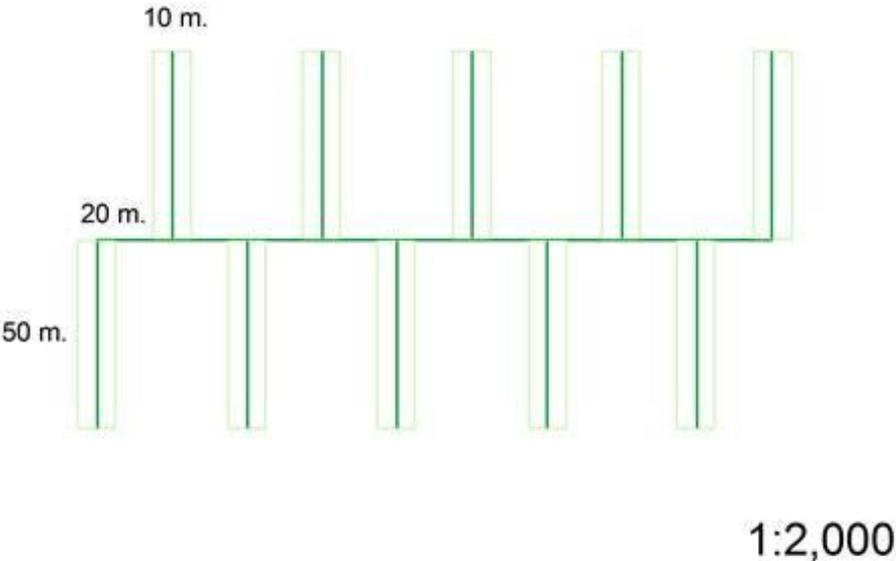
**b**



**c**



**Figure 3:** Diagram of the Gentry plot used in this study. The horizontal line was used as a reference to then establish transects (vertical lines) along which the inventory of trees was carried out. The distance between transects was 20m. Each transect was 10 m x 50 m.



**Table 1:** Coordinates of plot's location in Peruvian Amazonia. Units are in Universal Transverse Mercator coordinate system (UTM). Altitude of each plot above the sea level is also included. BV = Buena Vista, Q = Quistococha, SJ = San Jorge.

<b>Locality</b>	<b>UTM S</b>	<b>UTM W</b>	<b>Altitude</b>
BV 1	698576	9531660	115
BV 2	699246	9531720	103
BV 3	699739	9531496	111
BV 4	699303	9531160	96
BV 5	699411	9530617	104
BV 6	700051	9530803	102
BV 7	699738	9530073	101
BV 8	699925	9529551	103
Q 1	686475	9576239	109
Q 2	686841	9575885	103
Q 3	686355	9575608	102
Q 4	686766	9575059	111
Q 5	685995	9575029	96
Q 6	686202	9574381	115
Q 7	687128	9576929	100
Q 8	687156	9576313	96
SJ 1	701852	9551501	127
SJ 2	701609	9551134	128
SJ 3	701006	9550852	109
SJ 4	700330	9550793	120
SJ 5	699773	9550779	135
SJ 6	699212	9550755	125
SJ 7	701189	9550357	120
SJ 8	702217	9551087	117

**Table 2:** Plot scores for NMDS1 and NMDS2 that were extracted from the Non-metric multidimensional scaling ordination analysis (NMDS). BV = Buena Vista, Q = Quistococha, SJ = San Jorge.

<b>Peatland plot</b>	<b>NMDS1</b>	<b>NMDS2</b>
BV 1	-1.4217657	0.4926497
BV 2	-1.5754015	-0.0879906
BV 3	-1.7854119	-0.0063306
BV 4	-1.6411616	0.0865076
BV 5	-1.680151	0.1317459
BV 6	-1.9102245	0.0366095
BV 7	-1.8607512	0.2003836
BV 8	-1.7277039	0.3973680
Q 1	0.3900896	-0.8627702
Q 2	0.7129447	-0.5445510
Q 3	0.6414613	-0.8828875
Q 4	0.3147775	-0.4743073
Q 5	0.4290851	-0.7188143
Q 6	0.3633698	-0.5985037
Q 7	0.9434084	-0.5251703
Q 8	0.7685269	-0.6308620
SJ 1	0.810162	0.0987543
SJ 2	0.9859096	0.2390745
SJ 3	1.5456412	0.5483222
SJ 4	1.5084658	1.0443818
SJ 5	1.1145668	1.1551009
SJ 6	1.0434625	0.9003685
SJ 7	1.2210763	0.0570483
SJ 8	0.8096237	-0.0561274

**Table 3:** List of species per family. Number of individuals is shown per plot and study site.

Taxa	Buena Vista								Quistococha								San Jorge								Total	
	1	2	3	4	5	6	7	8	1	2	3	4	5	6	7	8	1	2	3	4	5	6	7	8		
<b>Anacardiaceae</b>																										
<i>Tapirira guianensis</i>										2	1									4	2				2	11
<b>Total</b>										2	1									4	2				2	11
<b>Annonaceae</b>																										
<i>Guatteria decurrens</i>													3	3	3				16	5	4	1		4	8	47
<i>Guatteria</i> sp. 1		2	2																							4
<i>Guatteria</i> sp. 2				1	2	1		2																		6
<i>Oxandra riedeliana</i>																			1	18	8	1		7		35
<i>Rollinia cuspidata</i>															1											1
<i>Unonopsis</i> sp. 1									1																	1
<b>Total</b>		2	3	2	1			2	1		1	3	3	3				16	6	22	9	1		11	8	94
<b>Apocynaceae</b>																										
<i>Aspidosperma schultesii</i>																								1		1
<i>Couma macrocarpa</i>																								1		1
<i>Himatanthus sucuba</i>									12		1	8	4		2	1	14	8			1	1	16	9	77	
<i>Lacmellea lactescens</i>									1				10													11
<i>Malouetia tamaquarina</i>	9	3		3			1	9	3		7	7	5	9		1								3	60	
<i>Parahancornia peruviana</i>									6	5			3	1		10									25	
<b>Total</b>	9	3		3			1	9	22	5	8	15	22	10	2	11	15	8			1	1	18	12	175	
<b>Aquifoliaceae</b>																										
<i>Ilex aff. nayana</i>		1	7	3	1	9	6	3																		30
<i>Ilex andarensis</i>									5	1													2	3	11	
<b>Total</b>		1	7	3	1	9	6	3	5	1													2	3	41	



<b>Chrysobalanaceae</b>										
<i>Licania heteromorpha</i>	3	10	2	5	3	2	2	1		28
<i>Licania licaniiflora</i>	1	4	1	1	2		2			11
<i>Licania octandra</i>								4		4
<i>Parinari parilis</i>		1		1	2	1	2			7
<b>Total</b>	<b>4</b>	<b>15</b>	<b>3</b>	<b>7</b>	<b>7</b>	<b>3</b>	<b>6</b>	<b>5</b>		<b>50</b>

<b>Clusiaceae</b>										
<i>Garcinia macrophylla</i>	8	10	5	2	4	2	1	4		36
<i>Symphonia globulifera</i>	1								17 17 14 19 19 11 10 9 7 3	10 7 144
<b>Total</b>	<b>9</b>	<b>10</b>	<b>5</b>	<b>2</b>	<b>4</b>	<b>2</b>	<b>1</b>	<b>4</b>	<b>17 17 14 19 19 11 10 9 7 3</b>	<b>10 7 180</b>

<b>Combretaceae</b>										
<i>Buchenavia amazonia</i>	8	3	9	3	17	8	4	16		68
<i>Buchenavia macrophylla</i>									1	1
<i>Terminalia dichotoma</i>	3	2	4	5	5		2	5		26
<b>Total</b>	<b>11</b>	<b>5</b>	<b>13</b>	<b>8</b>	<b>22</b>	<b>8</b>	<b>6</b>	<b>21</b>	<b>1</b>	<b>95</b>

<b>Ebenaceae</b>										
<i>Diospyros poeppigiana</i>	5	17	35	33	59	25	14	5		193
<b>Total</b>	<b>5</b>	<b>17</b>	<b>35</b>	<b>33</b>	<b>59</b>	<b>25</b>	<b>14</b>	<b>5</b>		<b>193</b>

<b>Elaeocarpaceae</b>										
<i>Sloanea cf. guianensis</i>								4	1	5
<i>Sloanea oppositifolia</i>	11	4	1		1		3	2		22
<b>Total</b>	<b>11</b>	<b>4</b>	<b>1</b>		<b>1</b>		<b>3</b>	<b>2</b>	<b>4</b>	<b>27</b>

<b>Euphorbiaceae</b>										
<i>Alchornea schomburgkii</i>								3	1 5 1 3 4 2 3 2 3	2 3 32

<i>Alchorneopsis floribunda</i>									1	2			1		1	3	2			1	2	1	2	2	18
<i>Conceveiba martiana</i>																	1	1						6	8
<i>Conceveiba rhytidocarpa</i>									1								3	2					9	22	37
<i>Hevea guianensis</i>									20		7	6	38	5	2	30	5			14		7		2	136
<i>Hura crepitans</i>									1															1	2
<i>Mabea cf. fistulifera</i>	1	2	4	3	4	4	1	3																	22
<i>Micrandra sp. 1</i>													18										1		19
<b>Total</b>	<b>1</b>	<b>2</b>	<b>4</b>	<b>3</b>	<b>4</b>	<b>4</b>	<b>1</b>	<b>4</b>	<b>25</b>	<b>2</b>	<b>8</b>	<b>11</b>	<b>58</b>	<b>8</b>	<b>7</b>	<b>35</b>	<b>14</b>	<b>5</b>	<b>3</b>	<b>15</b>	<b>2</b>	<b>8</b>	<b>14</b>	<b>36</b>	<b>274</b>

### Fabaceae

<i>Campsiandra angustifolia</i>								2	5																7	
<i>Crudia glaberrima</i>										1																1
<i>Cynometra bauhinifolia</i>																								1	3	
<i>Hydrochorea corymbosa</i>	4	17	8	15	13	18	13	8																	96	
<i>Inga cf. cayennensis</i>	2										2		1			2	3			4		2	1		17	
<i>Inga cf. psittacorum</i>		16	16	8	20	16	17	5																	98	
<i>Inga sp. 1</i>	1																								1	
<i>Inga sp. 2</i>			1	1	3		2																		7	
<i>Inga stenoptera</i>	7	15	51	25	36	64	29	35				1										1			264	
<i>Machaerium floribundum</i>											2				5		12	14	9				9	12	63	
<i>Macrolobium acaciifolium</i>	4	3	6	6	13	5	7	5																	49	
<i>Macrolobium cf. multijugum</i>					2					4	3	1	1	4	3	1									19	
<i>Ormosia coccinea</i> var. <i>subsimplax</i>																			3	2	1		7		13	
<i>Ormosia macrocalyx</i>		1						2																	3	
<i>Parkia multijuga</i>															1										1	
<i>Parkia nitida</i>																							1		1	
<i>Parkia panurensis</i>													1												1	
<i>Pterocarpus amazonum</i>			2	4	6	7	1	4	5			1											1		31	
<i>Swartzia schunkei</i>					2	4	1	2																	9	
<i>Vatairea guianensis</i>	22	28	21	27	27	14	20	9	13	2	13	39	18	21	3	10	6	4					10		307	
<i>Zygia cauliflora</i>		1	1					1				1			1			6							11	







<b>Proteaceae</b>											
<i>Roupala</i> sp. 1									2	1	3
<b>Total</b>									2	1	3

<b>Rubiaceae</b>															
<i>Platycarpum</i> sp. nov.									24	102	150	172	172	620	
<i>Remijia ulei</i>									2	2				4	
Rubiaceae sp. 1														1	1
<b>Total</b>									26	104	150	172	172	1	625

<b>Salicaceae</b>															
<i>Xylosma</i> cf. <i>tessmannii</i>	1														2
<b>Total</b>	1														2

<b>Sapindaceae</b>															
<i>Talisia</i> sp. 1	2														3
<i>Talisia</i> sp. 2										1	1	2			4
<b>Total</b>	2									1	1	2			7

<b>Sapotaceae</b>															
<i>Chrysophyllum argenteum</i>	3	1		1	1										7
<i>Elaeoluma glabrescens</i>	3	26	18	10	20	23	9	8							117
<i>Pouteria cuspidata</i>			2						2	3	1	5	2	3	18
<i>Pouteria cuspidata</i> subsp. <i>robusta</i>		3		1		2									6
<i>Pouteria gomphiifolia</i>		14	10	11	14	19		6							74
<i>Pouteria plicata</i>								1							1
<b>Total</b>	6	44	30	23	35	44	9	16	2	3	1	5	2	3	223

<b>Simaroubaceae</b>															
<i>Simaba orinocensis</i>			1	1	2	1	1	1							7



**Table 4 A:** Families with the highest number of species across the study plots.

<b>Family</b>	<b>Number of species</b>
Fabaceae	23
Euphorbiaceae	8
Lauraceae	8
Arecaceae	7

**Table 4 B:** Genera with the highest number of species across the study plots.

<b>Family</b>	<b>Genera</b>	<b>Number of species</b>
Fabaceae	<i>Inga</i>	5
Meliaceae	<i>Trichilia</i>	4
Moraceae	<i>Ficus</i>	4
Sapotaceae	<i>Pouteria</i>	4

**Table 5 A:** Families with the highest number of individuals across the study plots.

<b>Families</b>	<b>Number of individuals</b>
Arecaceae	1944
Fabaceae	1004
Rubiaceae	625
Calophyllaceae	598

**Table 5 B:** Genera with the highest number of individuals across the study plots.

<b>Family</b>	<b>Genera</b>	<b>Number of individuals</b>
Arecaceae	<i>Mauritia</i>	1216
Rubiaceae	<i>Platycarpum</i>	620
Arecaceae	<i>Mauritiella</i>	615
Calophyllaceae	<i>Calophyllum</i>	598

**Table 5 C:** Species with the highest number of individuals across the study plots.

<b>Family</b>	<b>Species</b>	<b>Number of individuals</b>
Arecaceae	<i>Mauritia flexuosa</i>	1216
Rubiaceae	<i>Platycarpum</i> sp. nov.	620
Arecaceae	<i>Mauritiella armata</i>	615
Calophyllaceae	<i>Calophyllum brasiliense</i>	597

**Table 6:** Total number of stems, species, and above-ground biomass (AGB) values per peatland plots in mega-grams per hectare (Mg ha<sup>-1</sup>). BV = Buena Vista, Q = Quistococha, SJ = San Jorge.

Peatland	N. of plot	N. of stems	N. of species	AGB (Mg ha <sup>-1</sup> )
BV	1	244	40	85.92
BV	2	318	42	111.27
BV	3	306	37	81.38
BV	4	267	36	68.29
BV	5	404	41	124.80
BV	6	312	31	75.99
BV	7	233	34	68.04
BV	8	302	45	71.90
Q	1	273	32	137.23
Q	2	407	18	121.98
Q	3	280	22	105.48
Q	4	337	29	118.06
Q	5	258	29	119.47
Q	6	245	23	119.50
Q	7	298	17	90.42
Q	8	352	22	101.82
SJ	1	387	31	115.04
SJ	2	356	31	112.03
SJ	3	373	20	89.55
SJ	4	513	11	109.89
SJ	5	394	14	121.33
SJ	6	355	13	189.79
SJ	7	325	28	88.28
SJ	8	292	34	117.82

**Table 7:** ANOVA table summarizing results of the differences in AGB across species. *df* = Degrees of freedom, MS = Mean squares, SS = Sum of squares.

<b>Source</b>	<b>SS</b>	<b>df</b>	<b>MS</b>	<b>F</b>	<b>P</b>
<b>Treatment (species)</b>	207.4	97	2.1384	13.27	< 2e-16
<b>Error</b>	337.6	2095	0.1612		
<b>Total corrected</b>	545	2192			

**Table 8:** Variance and covariance from the equation for estimating AGB. D = diameter, H = height, WD = wood density.

		<b>Variance</b>	<b>Covariance</b>	
<b>Variables</b>	<b>WD</b>	0.0759	<b>WD,D</b>	-0.0228
	<b>D</b>	0.3348	<b>WD,H</b>	0.0168
	<b>H</b>	0.0855	<b>D,H</b>	0.2376

**Table 9:** Variance (var) and covariance (cov) from the equation for estimating AGB per species. D = diameter, H = height, WD = wood density.

Species	Individual s per species	var(WD)	var(D)	var(H)	cov (WD,D)	cov (WD,H)	cov (D,H)
<i>Aniba guianensis</i>	3	0.0029	0.0012	0.0011	-0.0015	-0.0036	0.0006
<i>Roupala</i> sp. 1	3	0.0051	0.0083	0.0040	-0.0092	0.0046	0.0029
<i>Ormosia macrocalyx</i>	3	0.0002	0.0127	0.0021	0.0028	-0.0010	-0.0094
<i>Cynometra bauhiniifolia</i>	3	0.0001	0.0055	0.0031	-0.0011	0.0002	0.0034
<i>Pseudobombax munguba</i>	4	0.0022	0.0930	0.0057	-0.0079	0.0019	0.0352
<i>Remijia ulei</i>	4	0.0013	0.0043	0.0004	0.0044	0.0005	0.0010
<i>Ficus krukovii</i>	4	0.0030	0.8192	0.0428	-0.0099	0.0024	0.2304
<i>Talisia</i> sp. 2	4	0.0010	0.1429	0.0690	0.0199	0.0054	0.1324
<i>Socratea exorrhiza</i>	4	0.1450	0.0053	0.0039	-0.0429	0.0382	-0.0053
<i>Guatteria</i> sp. 1	4	0.0182	0.1453	0.0029	-0.0312	-0.0093	0.0250
<i>Sloanea</i> cf. <i>guianensis</i>	4	0.0026	0.2078	0.0355	-0.0348	-0.0125	0.1642
<i>Licania octandra</i>	4	0.0001	0.1365	0.0228	0.0036	0.0017	0.0987
<i>Micrandra</i> sp. 1	4	0.0415	0.2199	0.0475	-0.1481	-0.0529	0.1980
<i>Nectandra</i> sp. 1	4	0.0050	0.3024	0.0611	0.0348	0.0059	0.2580
<i>Lacmellea lactescens</i>	6	0.0014	0.1734	0.0713	0.0056	0.0020	0.1831
<i>Batocarpus amazonicus</i>	6	0.0003	0.1867	0.0073	0.0029	0.0001	0.0486
<i>Protium klugii</i>	6	0.0020	0.0240	0.0134	0.0010	0.0003	-0.0036
<i>Pachira insignis</i>	6	0.0019	0.1911	0.0111	-0.0064	0.0027	0.0708
<i>Guatteria</i> sp. 2	6	0.0042	0.1357	0.0079	-0.0289	-0.0072	0.0548
<i>Cespedesia spathulata</i>	6	0.0007	0.0364	0.0017	0.0060	0.0000	0.0049
<i>Nectandra heterotricha</i>	6	0.0005	0.1703	0.0082	0.0015	0.0007	0.0574
<i>Chrysophyllum argenteum</i>	6	0.0009	0.0161	0.0200	-0.0038	-0.0012	0.0228
<i>Campsiandra angustifolia</i>	6	0.0006	0.0465	0.0193	0.0089	0.0062	0.0457
<i>Trichilia rubra</i>	6	0.0110	0.0040	0.0054	-0.0018	-0.0036	0.0056
<i>Iryanthera paraensis</i>	6	0.0005	0.0565	0.0172	0.0044	0.0040	0.0522

<i>Pouteria cuspidata subsp. robusta</i>	6	0.0004	0.0550	0.0154	-0.0013	0.0019	0.0350
<i>Parinari parilis</i>	7	0.0020	0.0400	0.0071	0.0084	-0.0023	-0.0017
<i>Inga sp. 2</i>	7	0.0035	0.0243	0.0126	0.0011	0.0062	0.0208
<i>Conceveiba martiana</i>	7	0.0013	0.0671	0.0080	0.0038	0.0024	0.0279
<i>Simaba orinocensis</i>	7	0.0023	0.0252	0.0106	0.0023	0.0005	0.0266
<i>Ocotea gracilis</i>	8	0.0015	0.0373	0.0176	0.0016	0.0067	0.0128
<i>Coccoloba sp. 1</i>	8	0.0130	0.0105	0.0293	0.0080	0.0114	-0.0151
<i>Licania licaniiiflora</i>	9	0.0005	0.0839	0.0373	0.0018	0.0018	0.0934
<i>Swartzia schunkei</i>	9	0.0017	0.0472	0.0333	0.0053	-0.0010	0.0300
<i>Tapirira guianensis</i>	9	0.0086	0.1034	0.0128	0.0055	-0.0091	0.0556
<i>Ilex andarensis</i>	10	0.0165	0.0648	0.0148	0.0185	0.0111	0.0443
<i>Ormosia coccinea var. subsimplex</i>	11	0.0012	0.2173	0.0050	0.0068	0.0010	0.0510
<i>Oenocarpus bataua</i>	11	0.0407	0.0241	0.0130	0.0284	0.0209	0.0257
<i>Zygia cauliflora</i>	11	0.0014	0.0220	0.0110	0.0008	-0.0002	0.0227
<i>Pachira aquatica</i>	12	0.0032	0.0849	0.0570	0.0133	0.0202	0.0947
<i>Astrocaryum jauari</i>	12	0.0436	0.0294	0.0284	-0.0123	0.0423	0.0128
<i>Discocarpus brasiliensis</i>	13	0.0006	0.0486	0.0125	0.0037	0.0015	0.0405
<i>Ficus cf. guianensis</i>	14	0.0047	0.1120	0.0099	0.0000	-0.0051	0.0277
<i>Conceveiba rhytidocarpa</i>	16	0.0025	0.0512	0.0078	-0.0027	-0.0015	0.0107
<i>Coussapoa trinervia</i>	16	0.0014	0.0349	0.0069	-0.0013	-0.0012	0.0186
<i>Calyptranthes macrophylla</i>	17	0.0048	0.0506	0.0124	0.0078	0.0041	0.0240
<i>Sloanea oppositifolia</i>	17	0.0012	0.0646	0.0131	0.0093	0.0028	0.0430
<i>Oxandra riedeliana</i>	17	0.0018	0.0110	0.0095	0.0015	0.0007	0.0077
<i>Inga cf. cayennensis</i>	17	0.0018	0.0733	0.0162	-0.0033	-0.0014	0.0517
<i>Pouteria cuspidata</i>	18	0.0006	0.0380	0.0138	-0.0026	-0.0026	0.0299
<i>Alchorneopsis floribunda</i>	18	0.0016	0.0505	0.0156	0.0087	0.0039	0.0241
<i>Euterpe catinga</i>	19	0.0253	0.0296	0.0090	-0.0137	0.0004	0.0075
<i>Parahancornia peruviana</i>	19	0.0015	0.1296	0.0150	0.0051	0.0023	0.0659
<i>Macrobium cf. multijugum</i>	19	0.0019	0.1909	0.0339	0.0207	0.0049	0.1308
<i>Caryocar glabrum</i>	19	0.0020	0.0926	0.0178	0.0066	0.0022	0.0714
<i>Ruptiliocarpum caracolito</i>	20	0.0045	0.0574	0.0117	-0.0077	0.0007	0.0252

<i>Lueheopsis hoehnei</i>	21	0.0042	0.2329	0.0248	0.0168	0.0057	0.1216
<i>Ilex</i> aff. <i>nayana</i>	21	0.0078	0.0861	0.0163	0.0097	0.0033	0.0531
<i>Cecropia engleriana</i>	21	0.0129	0.0511	0.0118	0.0078	0.0063	0.0381
<i>Mabea</i> cf. <i>fistulifera</i>	22	0.0010	0.0618	0.0208	-0.0001	-0.0012	0.0509
<i>Platycarpum</i> sp. nov.	22	0.0015	0.0759	0.0098	-0.0012	-0.0009	0.0232
<i>Licania heteromorpha</i>	23	0.0029	0.0341	0.0133	0.0016	0.0038	0.0190
<i>Euterpe precatoria</i>	24	0.0516	0.0354	0.0309	0.0233	0.0284	0.0449
<i>Garcinia macrophylla</i>	25	0.0009	0.0923	0.0224	-0.0043	0.0000	0.0660
<i>Cecropia latiloba</i>	26	0.0139	0.0622	0.0129	0.0008	-0.0013	0.0257
<i>Pterocarpus amazonum</i>	26	0.0033	0.0033	0.0178	0.0073	0.0055	0.0194
<i>Terminalia dichotoma</i>	26	0.0027	0.1644	0.0303	0.0213	0.0086	0.1194
Myrtaceae sp. 1	29	0.0016	0.0326	0.0259	0.0011	0.0002	0.0153
<i>Machaerium floribundum</i>	31	0.0028	0.2109	0.0268	-0.0046	0.0009	0.0710
<i>Alchornea schomburgkii</i>	32	0.0029	0.0988	0.0346	-0.0018	0.0040	0.0590
<i>Guatteria decurrens</i>	32	0.0030	0.0928	0.0119	0.0065	0.0026	0.0399
<i>Pouteria gomphiifolia</i>	32	0.0052	0.0665	0.0151	0.0072	0.0032	0.0451
<i>Buchenavia amazonia</i>	33	0.0030	0.0350	0.0234	0.0053	0.0008	0.0215
<i>Eschweilera parviflora</i>	34	0.0034	0.1079	0.0257	0.0061	0.0031	0.0699
<i>Macrobium acaciifolium</i>	35	0.0028	0.1729	0.0405	0.0121	0.0063	0.1135
<i>Myrcia</i> sp. 1	35	0.0028	0.0607	0.0161	0.0093	0.0053	0.0463
<i>Inga</i> cf. <i>psittacorum</i>	35	0.0023	0.0590	0.0082	0.0044	0.0016	0.0231
<i>Dendropanax</i> cf. <i>resinosus</i>	37	0.0043	0.0168	0.0098	-0.0066	-0.0043	0.0026
<i>Brosimum utile</i>	38	0.0021	0.1964	0.0337	0.0052	0.0043	0.1127
<i>Elaeoluma glabrescens</i>	39	0.0083	0.0952	0.0261	-0.0114	0.0009	0.0807
<i>Pachira</i> aff. <i>brevipes</i>	39	0.0036	0.0831	0.0091	0.0107	0.0040	0.0351
<i>Hydrochorea corymbosa</i>	40	0.0039	0.0917	0.0169	0.0018	0.0031	0.0499
<i>Calophyllum brasiliense</i>	41	0.0036	0.0535	0.0103	0.0051	0.0019	0.0300
<i>Inga stenoptera</i>	41	0.0070	0.0459	0.0117	0.0094	0.0029	0.0161
<i>Triplaris americana</i>	41	0.0023	0.0564	0.0514	-0.0028	0.0005	0.0352
<i>Malouetia tamaquarina</i>	42	0.0156	0.0114	0.0094	-0.0001	0.0037	0.0081
<i>Diospyros poeppigiana</i>	42	0.0037	0.0577	0.0143	0.0013	0.0018	0.0259

<i>Eschweilera albiflora</i>	43	0.0026	0.1271	0.0269	0.0100	0.0051	0.0719
<i>Himatanthus sucuuba</i>	43	0.0028	0.0663	0.0267	-0.0006	0.0026	0.0374
<i>Tabebuia insignis</i> var. <i>monophylla</i>	50	0.0033	0.0445	0.0189	0.0020	0.0006	0.0241
<i>Hevea guianensis</i>	52	0.0116	0.0764	0.0266	0.0052	0.0093	0.0583
<i>Symphonia globulifera</i>	57	0.0018	0.1037	0.0220	0.0013	-0.0003	0.0713
<i>Virola surinamensis</i>	60	0.0046	0.1468	0.0334	0.0102	0.0031	0.1096
<i>Mauritiella armata</i>	62	0.1182	0.0089	0.0274	0.0130	0.0191	0.0089
<i>Eugenia</i> sp. 1	65	0.0043	0.0385	0.0243	0.0105	0.0008	0.0213
<i>Amanoa</i> aff. <i>guianensis</i>	75	0.0018	0.0646	0.0172	0.0023	0.0007	0.0256
<i>Mauritia flexuosa</i>	90	0.0283	0.0252	0.0188	-0.0010	0.0139	0.0108
<i>Vatairea guianensis</i>	90	0.0044	0.1215	0.0203	0.0022	0.0022	0.0639

---

**Table 10:** Spatial autocorrelation test “Moran I test” applied to regression residuals of the regression models in this study. Moran = correlation value.

<b>Hypothesis</b>	<b>Linear regression model</b>	<b>Moran</b>	<b><i>p</i></b>
Floristic composition	1. $TotalAGB_{plot} = a + b NMDS1_{plot}$	0.08	0.1
Intra-specific variation	6. $TotalAGB_{plot} = a + b \overline{AGBdeviation}_{plot}$	0.2	0.02
Floristic composition and Intra-specific variation	7. $TotalAGB_{plot} = a + b NMDS1_{plot} + \overline{AGBdeviation}_{plot}$	- 0.08	0.1

**Table 11:** Total AGB per plot ( $\text{Mg ha}^{-1}$ ) in other studies in western Amazonia. Sites: All = Allpahuayo Mishana, Jen = Jenaro Herrera, Suc = Sucusari, Yan = Yanamono. Forest types: FPF = Flooded palm forest, NPF = Non-flooded palm forest, PFPP = Partially-flooded palm forest, SF = Seasonally flooded, TF= Terra firme, TFC = Terra firme on clay soils, TFS = Terra firme on sandy soils.

<b>Author</b>	<b>Site</b>	<b>Area</b>	<b>Habitat</b>	<b>AGB</b>
Baker et al., 2004	All 11	0.44	TFC	248.43
Baker et al., 2004	All 12	0.4	TFS	202.4
Baker et al., 2004	All 21	0.48	TFS	232.27
Baker et al., 2004	All 22	0.44	TFC	186.5
Baker et al., 2004	All 30	1	TFS	187.58
Baker et al., 2004	Suc 01	1	TF	215.25
Baker et al., 2004	Suc 02	1	TF	221.66
Baker et al., 2004	Suc 03	1	SF	234.14
Baker et al., 2004	Suc 04	1	TF	241.06
Baker et al., 2004	Suc 05	1	TF	226.83
Baker et al., 2004	Yan 01	1	TF	245.44
Baker et al., 2004	Yan 02	1	TF	243.53
Honorio et al., 2009	13	0.5	NPF	159.1
Honorio et al., 2009	20 de Enero 01	0.5	PFPP	199.6
Honorio et al., 2009	20 de Enero 02	0.5	FPF	72.7
Honorio et al., 2009	Jen 14	0.5	FPF	183.7
Honorio et al., 2009	Jen 15	0.5	NPF	184.4

**Table 12:** Species scores for NMDS1 that were extracted from the Non-metric multidimensional scaling ordination analysis (NMDS).

<b>Species</b>	<b>NMDS1</b>
<i>Alchornea schomburgkii</i>	0.874158839
<i>Alchorneopsis floribunda</i>	1.071249158
<i>Amanoa</i> aff. <i>guianensis</i>	-0.365870307
<i>Aniba guianensis</i>	-1.905729673
<i>Aniba panurensis</i>	-1.720081454
<i>Aspidosperma schultesii</i>	1.3375881
<i>Astrocaryum jauari</i>	-1.787442977
<i>Batocarpus amazonicus</i>	0.885796946
<i>Brosimum utile</i>	0.917449573
<i>Buchenavia amazonia</i>	-1.877932496
<i>Buchenavia macrophylla</i>	0.391216166
<i>Byrsonima arthropoda</i>	-0.405081027
<i>Calophyllum brasiliense</i>	1.337274543
<i>Calophyllum</i> sp. 1	-1.885820598
<i>Calyptranthes macrophylla</i>	0.000661979
<i>Campsiandra angustifolia</i>	-1.95802567
<i>Caryocar glabrum</i>	-1.779433074
<i>Cecropia engleriana</i>	0.662112231
<i>Cecropia latiloba</i>	-1.383054441
<i>Cespedesia spathulata</i>	1.164307678
<i>Chrysobalanaceae</i> sp. 1	-1.551134951
<i>Chrysophyllum argenteum</i>	-1.732868481
<i>Coccoloba</i> sp. 1	-1.790632751
<i>Conceveiba martiana</i>	0.930229485
<i>Conceveiba rhytidocarpa</i>	0.887289041
<i>Couma macrocarpa</i>	1.3375881
<i>Coussapoa orthoneura</i>	0.339700897
<i>Coussapoa trinervia</i>	-1.655230433
<i>Crudia glaberrima</i>	0.42337491
<i>Cynometra bauhiniifolia</i>	-0.582678455
<i>Dendropanax</i> cf. <i>resinosus</i>	1.303072412
<i>Didymocistus chrysadenius</i>	0.445925303
<i>Diospyros poeppigiana</i>	-1.88933518
<i>Discocarpus brasiliensis</i>	-1.95704535
<i>Discophora guianensis</i>	0.882699688
<i>Elaeoluma glabrescens</i>	-1.894324559
<i>Eschweilera albiflora</i>	-1.897985541
<i>Eschweilera parviflora</i>	-1.754573534
<i>Eugenia</i> sp. 1	-1.283967848

<i>Euterpe catinga</i>	1.17004823
<i>Euterpe precatoria</i>	0.642982781
<i>Ficus cf. guianensis</i>	0.126397129
<i>Ficus krukovii</i>	0.5872054
<i>Ficus paraensis</i>	1.693323381
<i>Ficus trigonata</i>	1.026917676
Flacourtiaceae sp. 1	-1.792550922
<i>Garcinia macrophylla</i>	-1.832376831
<i>Guatteria decurrens</i>	1.033353161
<i>Guatteria</i> sp. 1	-1.85007236
<i>Guatteria</i> sp. 2	-1.859750148
<i>Heisteria spruceana</i>	-1.885820598
<i>Hevea guianensis</i>	0.937537509
<i>Himatanthus sukuuba</i>	0.902567977
<i>Hura crepitans</i>	0.645236734
<i>Hydrochorea corymbosa</i>	-1.891200654
<i>Ilex aff. nayana</i>	-1.953818904
<i>Ilex andarensis</i>	0.833571519
Indet sp. 1	0.067853523
Indet sp. 2	-1.720081454
Indet sp. 3	-1.720081454
Indet sp. 4	-1.836000777
Indet sp. 5	-1.551134951
<i>Inga cf. cayennensis</i>	0.82543981
<i>Inga cf. psittacorum</i>	-1.921911796
<i>Inga</i> sp. 1	-1.551134951
<i>Inga</i> sp. 2	-1.914136309
<i>Inga stenoptera</i>	-1.595782125
<i>Iryanthera paraensis</i>	0.42337491
<i>Iryanthera tessmannii</i>	-1.720081454
<i>Lacmellea lactescens</i>	0.456047323
<i>Leonia crassa</i>	0.42337491
<i>Licania heteromorpha</i>	-1.845507808
<i>Licania licaniiiflora</i>	-1.833388917
<i>Licania octandra</i>	-1.885820598
<i>Lueheopsis hoehnei</i>	0.723832977
<i>Mabea cf. fistulifera</i>	-1.889049785
<i>Machaerium floribundum</i>	1.129717869
<i>Macrobium acaciifolium</i>	-1.883491172
<i>Macrobium cf. multijugum</i>	0.563636928
<i>Malouetia tamaquarina</i>	-0.152216513
<i>Margaritaria nobilis</i>	-1.720081454
<i>Mauritia flexuosa</i>	0.986169471

<i>Mauritiella armata</i>	0.901428323
<i>Micrandra</i> sp. 1	0.645178886
<i>Myrcia</i> sp. 1	-1.886677404
Myrtaceae sp. 1	-1.899788788
<i>Nectandra heterotricha</i>	-1.648594378
<i>Nectandra</i> sp. 1	-0.635726158
<i>Nectandra</i> sp. 2	-1.951399088
<i>Neea</i> cf. <i>spruceana</i>	-1.551134951
<i>Ocotea aciphylla</i>	1.074981875
<i>Ocotea gracilis</i>	0.68201037
<i>Ocotea</i> sp. 1	1.228023428
<i>Oenocarpus bataua</i>	1.339512415
<i>Ormosia coccinea</i> var. <i>subsimplex</i>	1.504049249
<i>Ormosia macrocalyx</i>	-1.863851002
<i>Oxandra riedeliana</i>	1.549634113
<i>Pachira</i> aff. <i>brevipes</i>	1.402542014
<i>Pachira aquatica</i>	0.404694352
<i>Pachira insignis</i>	0.935160082
<i>Parahancornia peruviana</i>	0.660668168
<i>Parinari parilis</i>	-1.917698355
<i>Parkia multijuga</i>	1.026917676
<i>Parkia nitida</i>	0.878851117
<i>Parkia panurensis</i>	0.464606493
<i>Platycarpum</i> sp. nov.	1.374089968
<i>Pouteria cuspidata</i>	0.208834761
<i>Pouteria cuspidata</i> subsp. <i>robusta</i>	-1.091362778
<i>Pouteria gomphiifolia</i>	-1.904415786
<i>Pouteria plicata</i>	-1.885820598
<i>Protium klugii</i>	0.870427905
<i>Pseudobombax munguba</i>	-1.902270638
<i>Pterocarpus amazonum</i>	-1.051192282
<i>Remijia ulei</i>	1.448736806
<i>Rollinia cuspidata</i>	0.695912371
<i>Roupala</i> sp. 1	-1.962457772
Rubiaceae sp. 1	0.878851117
<i>Ruptiliocarpus caracolito</i>	1.061423353
<i>Simaba orinocensis</i>	-1.934935511
<i>Sloanea</i> cf. <i>guianensis</i>	0.438893945
<i>Sloanea oppositifolia</i>	-1.794560315
<i>Socratea exorrhiza</i>	0.980282135
<i>Styrax guianensis</i>	-2.037179741
<i>Swartzia schunkei</i>	-1.987498933
<i>Symphonia globulifera</i>	0.710078984

<i>Tabebuia insignis</i> var. <i>monophylla</i>	0.807014752
<i>Talisia</i> sp. 1	-1.687795027
<i>Talisia</i> sp. 2	0.523709413
<i>Tapirira guianensis</i>	1.067260682
<i>Terminalia dichotoma</i>	-1.836921821
<i>Trichilia inaequilatera</i>	-1.885820598
<i>Trichilia pleeana</i>	0.42337491
<i>Trichilia rubra</i>	-1.802437751
<i>Trichilia</i> sp. 1	0.391216166
<i>Triplaris americana</i>	-1.884295966
<i>Unonopsis</i> sp. 1	0.42337491
<i>Vatairea guianensis</i>	-0.318776196
<i>Virola surinamensis</i>	0.558266869
<i>Vismia angusta</i>	-1.551134951
<i>Xylosma</i> cf. <i>tessmannii</i>	-1.835305955
<i>Zygia cauliflora</i>	0.174551901
<i>Zygia longifolia</i>	-1.858711328
<i>Zygia</i> sp. 1	0.464606493

---

**Table 13:** Proportion of *M. flexuosa* in relation to its number of stems and to its total AGB at each plot. BV = Buena Vista, Q = Quistococha, SJ = San Jorge.

<b>Plot</b>	<b>Total AGB</b>	<b>Number of stems</b>	<b>Proportion of <i>M. flexuosa</i></b>	<b>Total AGB of <i>M. flexuosa</i></b>	<b>Proportion of the AGB of <i>M. flexuosa</i></b>
BV1	85.92	244	0	0	0
BV2	111.27	318	0	0	0
BV3	81.38	306	0	0	0
BV4	68.29	267	0	0	0
BV5	124.8	404	0	0	0
BV6	75.99	312	0	0	0
BV7	68.04	233	0	0	0
BV8	71.9	302	0	0	0
Q1	137.23	273	19.41	56.66	41.29
Q2	121.98	407	23.34	66.23	54.30
Q3	105.48	280	46.43	72.02	68.28
Q4	118.06	337	17.51	56.23	47.63
Q5	119.47	258	23.26	53.92	45.13
Q6	119.5	245	48.98	77.23	64.63
Q7	90.42	298	31.54	62.86	69.52
Q8	101.82	352	22.44	62.20	61.09
SJ1	115.04	387	27.91	66.97	58.22
SJ2	112.03	356	25.56	52.37	46.74
SJ3	89.55	373	12.87	23.12	25.82
SJ4	109.89	513	10.53	24.51	22.31
SJ5	121.33	394	5.08	11.04	9.10
SJ6	189.79	355	7.04	38.98	20.54
SJ7	88.28	325	34.77	56.73	64.26
SJ8	117.82	292	22.95	48.22	40.93